



DR. OLIVIER LAROCHE (Orcid ID : 0000-0003-0755-0083)

MR. OLIVER KERSTEN (Orcid ID : 0000-0001-9304-2493)

DR. ERICA GOETZE (Orcid ID : 0000-0002-7273-4359)

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Environmental DNA surveys detect distinct metazoan communities across abyssal plains and seamounts in the western Clarion Clipperton Zone

Running title: eDNA surveys of the abyssal CCZ

Olivier Laroche^{1*}, Oliver Kersten², Craig R. Smith¹, Erica Goetze¹

¹ *Department of Oceanography, School of Ocean and Earth Science and Technology, University of Hawaii at Manoa, Honolulu, USA;*

² *Centre for Ecological and Evolutionary Synthesis. University of Oslo, Norway;*

* Corresponding author,

Current address:

Institute of Marine Research,

PO Box 6606 Langnes

9296 Tromsø, Norway

E-mail: olli.laroche@gmail.com

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1 **Abstract**

2 The deep seafloor serves as a reservoir of biodiversity in the global ocean, with > 80 % of
3 invertebrates at abyssal depths still undescribed. These diverse and remote deep-sea communities are
4 critically under-sampled and increasingly threatened by anthropogenic impacts, including future
5 polymetallic nodule mining. Using a multi-gene eDNA metabarcoding approach, we characterized
6 metazoan communities sampled from sediments, polymetallic nodules, and seawater in the western
7 Clarion Clipperton Zone (CCZ) to test the hypotheses that deep seamounts (1) are species richness
8 hotspots in the abyss, (2) have structurally distinct communities in comparison to other deep-sea
9 habitats, and (3) that seafloor POC flux and polymetallic nodule density are positively correlated with
10 metazoan diversity. eDNA metabarcoding was effective at characterizing distinct biotas known to
11 occur in association with different abyssal substrate types (e.g., nodule- and sediment-specific fauna),
12 with distinct community composition and few taxa shared across substrates. Seamount faunas had
13 higher overall taxonomic richness, and different community composition and biogeography than
14 adjacent abyssal plains, with seamount communities displaying less connectivity between regions
15 than comparable assemblages on the abyssal plains. Across an estimated gradient of low to moderate
16 POC flux, we find lowest taxon richness at the lowest POC flux, as well as an effect of nodule size on
17 community composition. Our results suggest that while abyssal seamounts are important reservoirs of
18 metazoan diversity in the CCZ, given limited taxonomic overlap between seamount and plains fauna,
19 conservation of seamount assemblages will be insufficient to protect biodiversity and ecosystem
20 function in regions targeted for mining.

21 **1 | Introduction**

22 The deep seafloor serves as a reservoir of biodiversity in the global ocean, with > 80 % of
23 invertebrates at abyssal depths still undescribed (Smith, De Leo, Bernardino, Sweetman, & Martinez
24 Arbizu, 2008; Snelgrove & Smith, 2002). The vast and remote abyssal plains remain largely
25 unexplored (< 0.01 % sampled, Ramirez-Llodra et al., 2010), although they represent the dominant
26 topographical feature of the ocean seafloor (~ 70 %). Abyssal plains experience high physical stability
27 and are predominantly covered by fine sediments, providing habitat for diverse benthic communities
28 (e.g. Glover & Smith, 2003; Hannides & Smith, 2003; Smith et al., 2008). This demersal fauna
29 encounters a limiting allochthonous food supply and is characterized by slow growth, recruitment,
30 reproduction and recovery rates following disturbance (Huvenne, Bett, Masson, Le Bas, & Wheeler,
31 2016; Ramirez-Llodra et al., 2010).

32 Abyssal plains are punctuated by a multitude of seamounts (> 1000 meters above bottom
33 (mab); Harris, Macmillan-Lawler, Rupp & Baker, 2014) that may serve as hotspots for biodiversity
34 and potential refugia for populations impacted by environmental disturbances (Clark et al., 2010;
35 Rowden et al., 2010a). Seamounts are subject to distinct hydrodynamic processes and physical
36 conditions, including altered current velocity and organic matter deposition (White et al. 2008, Clark
37 et al. 2010). They have also been shown in some cases to support higher abundance and biomass of
38 benthic invertebrates than adjacent continental slopes (Beckmann & Mohn, 2002; Rogers, 1994;
39 Rowden et al., 2010a), and to serve as stepping stones for dispersal (Cho & Shank, 2010; Leal &
40 Bouchet, 1991; O'Hara, Consalvdey, Lavrado, & Stocks, 2010). Several emerging paradigms in
41 seamount ecology have not been fully tested or contradictory evidence has been found, including the
42 hypotheses that seamounts serve as species-richness hotspots, and that they have distinct species
43 composition or community structure, in comparison to adjacent deep seafloor habitats (Rowden,
44 Dower et al., 2010, McClain, Lundsten, Ream, Barry & DeVogelaere, 2009). Seamounts have also
45 been hypothesized to function as biogeographic 'islands', harboring high levels of endemism (Koslow
46 et al., 2001; McClain, Lundsten et al., 2009; Samadi et al., 2006; Stocks & Hart, 2007; Wilson and
47 Kaufman, 1987), yet a number of studies have reported low levels of seamount endemism with

48 greater sampling effort (Hall-Spencer, Rogers, Davies, & Foggo, 2007; Samadi et al., 2006). Most
49 seamounts studied to date have bathyal or shallower summit depths and occur in proximity to
50 continental slopes; little is known about abyssal seamounts in remote areas of the central Pacific.

51 The Clarion Clipperton Zone (CCZ) deep-sea floor holds significant metal and mineral
52 resources in the form of polymetallic nodules (Ramirez-Llodra et al., 2010). With dwindling onshore
53 mineral reserves and security concerns over supply, there is renewed interest in mining the deep
54 seafloor, as shown by a tripling in the number of exploration mining claims granted by the
55 International Seabed Authority (ISA) in the past eight years (Fukushima et al., 2017). The CCZ holds
56 the highest abundance of polymetallic nodules of commercial interest of any region in the global
57 ocean, with 16 of the 18 active nodule exploration contracts granted by the ISA within the CCZ
58 (Wedding et al., 2015; Wedding et al., 2013; ISA website <https://www.isa.org.jm>). The ISA has
59 designated nine no-mining areas, termed Areas of Particular Environmental Interest (APEI), each
60 160,000 km², to safeguard regional biodiversity in the face of nodule mining (Wedding et al., 2013).
61 The APEIs span large-scale physical and biological gradients (Wedding et al., 2015; Wedding et al.,
62 2013), but there is limited ecological information available from APEIs, hindering accurate
63 assessment of their regional representativity (De Smet et al., 2017; Gollner et al., 2017; Miller et al.,
64 2019). Fundamental ecological knowledge, including levels of biodiversity, community composition,
65 species ranges and population connectivity among habitats in these regions remain largely unknown
66 (Kaiser, Smith, & Arbizu, 2017).

67 Polymetallic nodules represent an important structuring element within the CCZ seafloor
68 habitat, providing hard substrate microhabitats within the extensive soft sediments of the abyssal
69 plains. Nodules support sessile organisms, such as xenophyophores, antipatharian corals and sponges,
70 as well as numerous other megafaunal, meiofaunal and microbial taxa (Amon et al., 2016; Shulse et
71 al., 2017; Thiel et al. 1993; Vanreusel et al. 2016; Veillette et al., 2007). Nodules influence the
72 community composition and distribution of abyssal biota, and positively affect organismal abundance
73 and diversity (e.g. Mullineaux, 1987; Shulse et al., 2017; Vanreusel et al. 2016; Veillette et al. 2007).
74 Nodule mining will remove and bury the nodule, hard-substrate habitats and cause resuspension of the
75 upper ~ 5 cm sediment layer (Oebius et al., 2001; Thiel et al., 2001); thus, nodule mining is expected
76 to have substantial disturbance effects on benthic communities (Glover & Smith, 2003; Jones et al.,

77 2018). Seamounts within the CCZ might harbour refugial populations and provide larval sources for
78 the hard-substrate biota likely to be obliterated by large-scale mining operations on the abyssal plains,
79 but they remain almost entirely unstudied.

80 Environmental DNA (eDNA) metabarcoding surveys can provide baseline assessments of
81 biodiversity that may circumvent some of the challenges of comprehensively sampling remote and
82 highly diverse communities in deep ocean habitats (Boschen et al., 2016). Methods based on eDNA,
83 herein defined to include both intra- and extra-cellular DNA, are particularly informative for
84 detecting rare, cryptic and invasive species (Cristescu & Hebert, 2018, Kersten et al. 2019). Many
85 species in the abyssal CCZ are undescribed (e.g. Amon et al., 2016; Tilot et al., 2018), and whole
86 community sequencing could provide a valuable baseline community assessment prior to mining,
87 with limited dependence on taxonomic species descriptions. eDNA metabarcoding is increasingly
88 being used to characterise and monitor marine ecosystems (Danovaro et al., 2016; Everett & Park,
89 2018; Goodwin et al., 2017), but has seen limited application in the deep sea. Recent eDNA studies
90 on deep ocean sediments have shown high local heterogeneity, and a high proportion of
91 uncharacterized species in eukaryotic communities (Lejzerowicz et al. 2014; Dell'Anno et al. 2015;
92 Guardiola et al. 2015, 2016; Sinniger et al. 2016).

93 Using a multi-gene eDNA metabarcoding approach, we aimed to comprehensively
94 characterize metazoan communities in three APEIs in the western CCZ (APEIs 1, 4 and 7), and test
95 several hypotheses regarding diversity across environmental gradients in the abyssal benthos. First,
96 we compare community composition and diversity between samples from three different substrates,
97 seafloor sediments, polymetallic nodules, and seawater from the benthic boundary layer (BBL), to
98 evaluate how effectively eDNA metabarcoding distinguishes the distinct biotas known to occur in
99 these different substrate types (e.g. Amon et al. 2016, Simon-Lledó et al., 2019). Then we test the
100 hypotheses that deep seamounts (1) are species richness hotspots in the abyss, (2) have distinct
101 community composition and biogeography in comparison to other deep sea habitats, and (3) that
102 seafloor particulate organic carbon (POC) flux and polymetallic nodule density are positively
103 correlated with metazoan diversity. We discuss our results in the context of future deep seabed mining
104 and the potential importance of biodiversity hotspots to conservation of metazoan communities at the
105 abyssal seafloor.

106 2 | Materials and methods

107 2.1 | Field sampling

108 Samples from seafloor sediment, polymetallic nodules and seawater were collected in APEIs 1, 4 and
109 7 within the western CCZ between May 22 and June 12, 2018 aboard cruise 18-08 on the *RV Kilo*
110 *Moana* (DeepCCZ cruise), using the *ROV Lu'ukai* (Fig. 1). Sampling targeted one seamount and the
111 adjacent abyssal plain habitat within each APEI. Sampled seamounts were elongate features with
112 summit depths of 3100 m (APEI 7), 3500 m (APEI 4), and 3900m (APEI 1), all with summits >1000
113 m above the surrounding abyssal plain. Adjacent abyssal plain sites were sampled >15 km away from
114 the seamount ridgeline (APEI 7) or base (APEIs 4 and 1), with the expectation that this would be
115 outside the 'zone of influence' of the seamount, although limited data are available from the deep sea
116 with which to estimate the appropriate scale of seamount influence. Large seamounts with shallow
117 summit depths are relatively better studied, and for these features, seamount effects have been
118 documented to a radius of up to 30 km. In the deep ocean, current velocities are generally an order of
119 magnitude lower than in the energetic top 500 m of the water column; therefore, to be conservative,
120 we chose a 15 km buffer from the summit of the seamount to the nearest abyssal-plain sampling sites.
121 The seamount in APEI 1 was sampled for seawater only.

122 The *ROV Lu'ukai* was used to sample sediments and nodules, with three dives in APEI 7 (2
123 abyssal plain, 1 seamount), three dives in APEI 4 (2 abyssal plain, 1 seamount), and two abyssal plain
124 dives in APEI 1, with 2-5 sediment cores collected for eDNA on each ROV dive. Seven-centimeter
125 diameter push cores were gently inserted vertically into the sediment by the ROV, sealed and
126 recovered in the ROV work basket, and then horizontal sectioned on shipboard into 0-2 cm and 3-5
127 cm sediment intervals for eDNA. Sterile syringes (60 mL; single-use) were used to extract mini-cores
128 from each sediment interval. Sediment processing gear and push-core tubes were treated with 10 %
129 bleach and rinsed with ddH₂O between each ROV dive to prevent contamination. Slicing equipment
130 was rinsed in ddH₂O between cores. Two eDNA minicore technical replicates were taken for all cores
131 from APEI 1. Samples were cryopreserved at -80 °C until further processing. Polymetallic nodules
132 were either collected in push cores, or by the manipulator arm of the ROV and placed in a sealed

133 sample box (BioBox) for shipboard recovery. Once brought on shipboard, nodules were transferred to
134 sterile whirl-pack bags and cryopreserved (-80 °C). Supplementary file 2 lists all ROV push cores
135 sampled for eDNA.

136 Seawater samples were collected using conductivity-temperature-depth (CTD) casts with a
137 rosette sampler with 24 x 10L Niskin bottles (SBE 911plus/917plus, SeaBird oxygen sensor (SBE43),
138 Seapoint fluorometer, Wetlabs C-Star transmissometer). A total of 12 CTD casts were conducted
139 during the cruise, with two over the abyssal plain and two over the seamount within each APEI (Table
140 S1). Niskin bottles were collected at seven depths within the water column: 5 meters above bottom
141 (mab), 50 mab, bathypelagic depths (3000 m over plains, 2500 or 2000 m over seamounts), deep
142 mesopelagic at 1000 m, mesopelagic at 500 m, deep chlorophyll maximum (DCM; between 90 m to
143 60 m), and 5 m in the near sea surface. Seawater volumes filtered were variable across depth, 5 L per
144 replicate at 5 mab, 50 mab and bathypelagic depths, 4 L in the deep mesopelagic (1000m), 2 L in the
145 mesopelagic (500m), and 1 L at the DCM and in the near surface, with 4 to 6 replicates taken from
146 each cast and depth. Field negative controls (double-distilled water; ddH₂O) were collected for each
147 CTD cast, with filtration and handling as for all other bottles. Seawater was filtered onto 0.2 µm
148 sterile Supor filters (Pall) using 47 mm inline polycarbonate filter holders and two peristaltic pumps.
149 Filters were immediately preserved in 1 ml of RNALater (Invitrogen), flash frozen in liquid nitrogen,
150 and held at -80°C until further processing. During the sampling process, carboys, tubing, plastics and
151 the workspace were treated with 10 % bleach for a minimum of 30 minutes to minimize cross-
152 contamination, followed by three ddH₂O and three seawater rinses. To avoid contamination during
153 sample collection, personal protective equipment included disposable lab coats and nitrile gloves for
154 all involved personnel.

155 2.2 | Sample processing and library preparation

156 Environmental DNA was extracted from sediment samples using the PowerMax® Soil kit
157 (QIAGEN, California, USA) following the manufacturer's protocol. Approximately 10 g of
158 homogenized sediment (mixed with a sterile metallic spatula) was used per extraction. Captured and
159 purified DNA was eluted in 1 mL and then 4 mL ddH₂O. Polymetallic nodules were weighed, and
160 eDNA extraction performed by first grinding and homogenising nodules inside their whirl-pack bag

161 using a 16 g ceramic pestle. Ten subsamples of ~500 mg per nodule were used for eDNA extraction
162 with the FastDNA™ Spin kit according to the manufacturer's instructions. To obtain sufficient DNA
163 for polymerase chain reaction (PCR) amplification, subsamples were pooled in pairs (mean DNA
164 concentration of 0.382 ng/μL) and concentrated to ~ 1 ng/μL with the DNA Clean & Concentrator kit
165 (Zymo Research, California, USA), resulting in five replicates per nodule. Environmental DNA from
166 seawater samples was extracted with the DNeasy® Plant Mini kit (QIAGEN, California, USA), using
167 a modified protocol as described in Laroche et al. (2020). Due to low eDNA concentration in the 5
168 mab and 50 mab samples, two replicates per collection point (2 x 5 L of filtered seawater for each
169 depth) were pooled. For all sample types (seawater, sediment, nodules), an extraction blank was used
170 to assess potential contamination during sample processing. All sample handling and DNA extraction
171 steps were carried out in a dedicated laboratory free of PCR-amplified DNA.

172 Eukaryotic communities were characterized by amplicon sequencing using two genetic
173 markers, the V4 region of the 18S rRNA gene (approximately 450 base pairs [bp]) and a fragment (ca.
174 350 bp) of the mitochondrial cytochrome *c* oxidase subunit 1 (COI) gene. For 18S rRNA, the
175 eukaryotic forward Uni18SF: 5'-AGG GCA AKY CTG GTG CCA GC-3' and reverse primers
176 Uni18SR: 5'-GRC GGT ATC TRA TCG YCT T-3' primers (Zhan et al., 2013) were used. For COI,
177 amplifications used the universal metazoan primers mlCOIintF: 5'-GGW ACW GGW TGA ACW
178 GTW TAY CCY CC-3' and jgHCO2198: 5'-TAI ACY TCI GGR TGI CCR AAR AAY CA-3' (Leray
179 et al., 2013, Geller et al. 2013). Details regarding library preparation can be found in Supplementary
180 file 1. Unprocessed sequencing reads are available from the NCBI Sequence Read Archive (SRA)
181 under accession numbers SRR9199590 to SRR9199853.

182 2.3 | Bioinformatic analysis

183 Samples were demultiplexed by their 8-mer Nextera index, and then demultiplexed by target
184 gene using cutadapt (version 1.8; Martin, 2011). Sample reads were denoised with the DADA2
185 program (Callahan et al., 2016) implemented in Qiime2 (version 2018.11; Boylen et al., 2018) using
186 the default parameters. Denovo chimera detection was performed using the consensus approach.
187 Forward and reverse reads were truncated at 260 base pairs (bp) and 235 bp for 18S rRNA, and at
188 260, 215 for COI, respectively, and merged using a perfect minimum overlap of 20 bp. Trimming of

189 the 3' end of the forward and reverse reads was performed to reduce Phred-score based expected error
190 of the sequences, and increase the yield of good quality, denoised reads. For 18S rRNA, taxonomic
191 assignment for each read was performed with a naive Bayes classifier (Bokulich et al., 2018)
192 implemented in Qiime2 and trained on a trimmed SILVA 18S rRNA database (release 132 clustered
193 at 99 % similarity; Wang et al., 2007). For COI, taxonomic assignment was achieved using a
194 combination of approaches that included the use of the classification trees ('insect') classifier (version
195 5; Wilkinson et al., 2018), and megablast and blastn methods (Camacho et al., 2009) applied to the
196 GenBank nucleotide (nt) database (Benson, Karsch-mizrachi, Lipman, Ostell, & Wheeler, 2008).
197 Complete description of the methods used in taxonomic assignment can be found in Supplementary
198 data.

199 2.4 | Data analysis and statistics

200 Sequencing depth and recovered diversity per sample was investigated using rarefaction
201 curves with the 'vegan' R package (Oksanen et al., 2018). Prior to data analysis, sequences found in
202 all negative controls, including field (ddH₂O), DNA extraction and PCR blanks were investigated
203 (Table S2) and removed from the dataset. Sequences unidentified at kingdom level or not part of
204 Metazoa, and those originating from non-marine taxa were also discarded (72 % of 18S reads and 64
205 % of COI reads). For COI, two datasets were explored, one using amplicon sequence variants
206 (ASVs), and one of operational taxonomic units (OTUs) from ASVs clustered at 97 % similarity
207 using the default parameters of Vsearch (Rognes, Flouri, Nichols, Quince, & Mahé, 2016)
208 implemented in Qiime2.2018-11. The OTU-level analysis aims to achieve putative species-level
209 taxonomic resolution. To remove pelagic legacy eDNA, or eDNA that derives from organisms living
210 in overlying pelagic ecosystems, all ASVs found in the water column (from 5 m in the near sea
211 surface to 2,000 m [seamount] or 3,000 m [abyssal plain]) were discarded from the deep-sea samples
212 (sediment, nodules, 5 mab and 50 mab benthic boundary layer (BBL) seawater samples), as in
213 Laroche et al. (2020). To simplify analyses, sample data from both the 0-2 and 3-5 cm sediment
214 horizons were combined, representing eDNA collected from a total of 20 g of sediment per sample.
215 Taxonomic composition of the sediment, polymetallic nodules and BBL samples was visualized with
216 a cladogram containing a circular heatmap and barplots using GraPhlAn (Asnicar, Weingart, Tickle,

217 Huttenhower, & Segata, 2015) and the metacoder R package (Foster, Sharpton, & Grünwald, 2017).
218 For this analysis, only taxa found in a minimum of five samples were included. ASV and OTU
219 richness, estimated with the Chao2 index, was used to compare alpha-diversity between sample types,
220 APEIs, and habitats at base coverage. Base coverage is defined as the highest coverage value between
221 minimum extrapolated values and maximum interpolated values (*see* Chao et al. 2014), and we use it
222 as a metric for comparison among samples that standardizes for sampling coverage (or completeness).
223 Calculations were performed using the iNEXT R package (Hsieh, Ma, & Chao, 2016). Only sediment
224 samples were considered for the comparison between APEIs and habitats (seamount, plain). ASVs
225 and OTUs shared between sample types were investigated with Venn diagrams and the eulerr R
226 package (Larsson, 2019). To avoid any bias from sampling coverage, each sample type dataset was
227 subsampled at equivalent coverage (determined by the Chao2 index) with 50 iterations. Mean
228 metazoan and phyla richness per sample and sample source were visualized with stacked barplots,
229 plotted using the ggpubr R package (Kassambara, 2018).

230 Beta-diversity analysis was conducted using unweighted UniFrac distance matrices (Lozupone
231 & Knight, 2005) within phyloseq (McMurdie & Holmes, 2013), and visualized with non-metric
232 multidimensional scaling plots (nMDS). The matrices were based on phylogenetic trees produced in
233 Qiime2 using the phylogeny align-to-tree-mafft-fasttree command (Kato & Standley, 2013; Price,
234 Dehal, & Arkin, 2010) and default parameters. Homogeneity of variance within sample type, APEI
235 and habitat groups was analyzed with the betadisper function of the vegan package. Differences in
236 beta-diversity between sample types, APEIs and habitats were assessed with pairwise permutational
237 analysis of variance (permanova) in the vegan R package. The effect of nodule weight on community
238 composition was assessed by permanova using the adonis function of the vegan package, with nodule
239 weight nested within APEI. To correspond as closely as possible to traditional morpho-taxonomy
240 studies, both alpha and beta-diversity analyses were performed on the COI dataset clustered into
241 OTUs at 97 % similarity (putative species-level differentiation).

242 Using presence/absence data, the proportion of taxa either unique to each habitat (seamounts,
243 abyssal plains) and APEI, unique to a habitat but not to an APEI ('widespread habitat-specific') or
244 found in diverse habitats and APEIs ('widespread non-specific') was visualized at both the ASV level
245 (18S and COI) and OTU level (COI) using bar plots (plotted using the ggplot2 R package). For this

246 analysis, only sediment samples were considered. To account for uneven sampling among
247 APEI:Habitat combinations, the biogeography category assignment of each ASV/OTU was carried
248 out by randomly subsampling each APEI:Habitat group to the sample size of the smallest group (e.g.,
249 2 cores), and by performing 100 iterations. Differences in the proportion of unique, widespread
250 habitat-specific and widespread non-specific taxa among habitats were tested with a Kruskal-Wallis
251 rank sum test. The choice to use a non-parametric test was motivated by significant differences
252 observed in group variance based on a Levene's test.

253 **3 | Results**

254 **3.1 | High-throughput sequencing**

255 A total of 10,315,003 and 17,202,778 reads were generated for 18S and COI, respectively
256 (Table S3). Quality filtering, denoising, merging and chimera removal reduced 18S read counts by 54
257 % and COI read counts by 40 %, leaving an average of 20,040 and 43,572 good quality reads per
258 sample for 18S and COI, respectively. ASVs found in sampling and extraction blanks were removed
259 from all samples and are reported in Table S2. Rarefaction curves indicated that all but one sample
260 (18S N-26) were sufficiently sequenced to capture total amplicon within-sample richness (reached an
261 asymptote, Figs. S1, S2). This sample, along with two COI seawater samples with very few reads (<
262 3,000 reads; W-416-417, W-74-75) were excluded from all downstream analyses.

263 While only 7 % of 18S sequences could not be assigned to a domain, unclassified COI
264 sequences at the level of domain represented 59 % of reads. Once these unclassified reads were
265 removed, the proportion of sequences derived from Metazoa was 30 % for 18S and 90 % for COI.
266 Among sample types, seawater samples contained the lowest proportion of metazoan reads (20 %
267 [18S] and 78 % [COI]). Protists (SAR supergroup) corresponded to 69 % and 8 % of all reads,
268 respectively, while Fungi and Viridiplantae composed less than 1 % and 2 % of 18S and COI reads.
269 Keeping only metazoan taxa resulted in a total of 2,020 and 11,901 ASVs, and 1,308,427 and
270 2,802,156 reads for 18S and COI data, respectively. Removing ASVs found in the pelagic
271 environment reduced the 18S dataset to 1,759 ASVs and 839,626 reads, and the COI dataset to 9,574

272 ASVs and 2,333,545 reads. Clustering COI ASVs at 97 % similarity resulted in a total of 6,282 OTUs
273 sampled in the abyss (all sample types).

274 3.2 | eDNA taxonomic resolution

275 The level of taxonomic identification achieved varied substantially between marker genes,
276 with much higher proportions of 18S rRNA reads assigned taxonomy at Phylum to Species levels
277 (Table 1). For 18S, the phyla with the highest taxonomic resolution (ASVs identified to species level)
278 with a minimum of 10 ASVs were Xenacoelomorpha (80 %), Gastrotricha (73 %), Chordata (50 %),
279 and Bryozoa (50 %) (Table S4). Phyla with the lowest resolution included Nematoda (9 % of ASVs
280 identified to species), Ctenophora (9 %), Nemertea (0 %) and Loricifera (0 %; Table S4). For COI
281 data, the only phylum with high taxonomic resolution was Chordata, with 90 % of OTUs identified at
282 the species level (Table S4). Among the remaining most read-count dominant phyla, the percentage of
283 OTUs identified at Species and Genus levels (COI), respectively, were 5 % and 23 % for
284 Echinodermata, 6 % and 10 % for Mollusca, 3 % and 4 % for Porifera, 2 % and 6 % for Annelida, 2
285 % for Cnidaria and Arthropoda, and 0 % for Platyhelminthes and Nemertea (Table S4).

286 3.3 | Taxonomic composition and community diversity

287 3.3.1 | Sample type

288 Overall, a mean of 7, 17 and 51 unique 18S ASVs could be recovered per BBL seawater (10
289 L), polymetallic nodule (5 g) and sediment (20 g) sample (Fig. 2). Metazoan diversity resolved in the
290 18S rRNA data was composed of 19 Phyla, 35 Classes, 71 Orders and 97 Families, largely dominated
291 by nematodes (23 % ASVs), cnidarians (16 % ASVs), annelids (11 % ASVs), and arthropods (10 %
292 ASVs). In terms of reads (Fig. 3), nematode and arthropod (harpacticoid copepod) reads were
293 predominantly found in both sediments and on nodules, while annelid, cnidarian, bryozoan,
294 brachiopod, echinoderm, mollusk, and poriferan reads were mostly present on nodules. Reads
295 sampled in seawater mostly derived from cnidarians (narcomedusae, trachymedusae), ctenophores,
296 and arthropods (calanoids) (Fig. 3). Several taxa were found to be exclusive to a particular substrate
297 type. Considering taxa present in at least five samples, 79 ASVs were found to be exclusively present

298 on nodules (Table S5F), including brachiopods (Terebratulida), ascidians (Styelidae), corals
299 (Isididae), bivalves (Veneroidea, Mytiloidea), hydroids (Ptilocodiidae), bryozoans, sponges
300 (Cladorhizidae, Suberitida), turbellarian worms, polychaetes (Phyllodoceidae, Syllidae), and
301 scyphozoan cnidarians. Most of the 197 ASVs exclusive to sediments (≥ 5 samples) were nematodes
302 (Xyalidae, Comesomatidae, Enoplida), although 29 ASVs were classified as harpacticoids or
303 arthropods and 14 were catenulid flatworms. A range of hydrozoan cnidarian groups as well as
304 several other taxa were found to be exclusive to the BBL (e.g., Narcomedusae, Rhopalonematidae),
305 but had lower recurrence across samples (occurrence in < 5 samples; Table S5).

306 For COI, the mean number of recovered COI OTUs per sample was 30, 118 and 211 for BBL
307 seawater (10 L), nodules (5 g) and sediments (20 g) (Fig. 2). Overall, 19 Phyla, 29 Classes, 51 Orders,
308 and 55 Families could be identified. Most of the OTU richness could be taxonomically assigned only
309 to Metazoa (79 % OTUs), with the remainder mostly assigned to arthropods (8 % OTUs), cnidarians
310 (6 % OTUs), poriferans (3 % OTUs), annelids and molluscs (1% OTUs). Although large numbers of
311 reads and COI OTUs could not be taxonomically classified beyond Metazoa, their association to
312 sample type and distribution across habitats could be resolved within the scope of our data. Of the
313 ~30% of reads that could be assigned taxonomy to phylum or below, most cnidarian, annelid and
314 echinoderm reads were sampled on nodules, sediments contained arthropods and cnidarians, and reads
315 in BBL seawater samples were dominated by arthropods, cnidarians, poriferans, echinoderms and
316 chordates (Fig. S3). The proportion of unclassified metazoan ASVs was highest within nodule
317 samples (82 %), followed by sediment (77 %) and BBL (56 %) samples. Taxa that were restricted to a
318 particular substrate type and present in at least 5 samples included sponges, such as hexactinellids and
319 suberitids, for nodules (25 OTUs), and Chromadorea (nematodes) for sediments (4 OTUs). Cetacea,
320 Scombriformes and hydrozoan siphonophores, including Apolemiidae, Diphyidae, Forskaliidae, and
321 Sphaeronectidae, were found exclusively in BBL seawater, but had lower recurrence across samples
322 in some cases (occurrence < 5 samples; Fig S3, Table S6).

323 Despite relatively low sampling coverage of ASV and OTU richness in sediment (37 % and 60
324 %, respectively) and seawater samples (41 % and 50 %, respectively), Figure 4 shows that at base
325 coverage, or the highest coverage value between minimum extrapolated values and maximum
326 interpolated values (Chao et al. 2014), sediments contained from 2.5 (COI) to 12.6 (18S) times the

327 richness of BBL seawater or nodules. A significant difference can also be observed between seawater
328 and nodules, but for COI data only, the latter containing twice as many estimated OTUs as seawater
329 (Fig. 4).

330 Community composition differed significantly between sample types for both target genes,
331 with stronger grouping by sample type within the COI data (Fig. 5A-B). Pairwise permutational
332 analysis of variance showed strongest dissimilarity between water samples and sediment or nodule
333 samples for both target genes (Table S7). The analysis of homogeneity of variance among sample
334 types was also significant ($p < 0.043$ both markers; Table S8), possibly due to the effect of habitat
335 (seamount, plain).

336 Using a normalised approach in which the numbers and proportions of shared ASVs (18S) and
337 OTUs (COI) between sample types were analysed at equivalent sampling coverage (40 and 50 % for
338 18S and COI, respectively), in order to control for sampling effort, our analyses showed very little
339 sequence overlap among substrates (Fig. 5C-D). The highest proportion of shared sequences was
340 found between sediment and nodules (mean of 1.2 and 4.2 % of all ASVs and OTUs at equivalent
341 coverage for 18S and COI, respectively; Fig. 5C-D). <1 % of BBL ASVs and OTUs were found
342 within sediment and nodule samples (Fig. 5C-D).

343

344 3.3.2 | APEI

345 At the same sampling coverage, taxon richness tended to be slightly higher within APEI4 than
346 APEI7, and lowest in APEI1 (Fig. 4). Community composition was significantly different between
347 APEIs for all sample types and target genes except 18S BBL seawater samples ($p < 0.01$; Table S9).
348 Additionally, individual nodule weight significantly affected community composition, and to a greater
349 extent than APEIs ($p = 0.001$; Table S9). Differences in community composition between APEIs were
350 more pronounced in the COI data, where pairwise analysis found significant differences between all
351 APEI combinations and for each sample type ($p < 0.02$; Table S10). In contrast, significant
352 differences in community composition between APEIs in the 18S data were found only for nodules
353 ($R^2=0.088$, $p = 0.001$, Table S9). Overall, the level of community dissimilarity between the different
354 APEI pairwise comparisons were relatively similar (R^2 from 0.04 to 0.11; Table S10), with no clear
355 association with geographic distance. A Mantel test using spatial coordinates and biological

356 community dissimilarity matrices found significant correlations for nodules ($p < 0.001$, Table S11)
357 and for COI sediment samples ($p = 0.002$, Table S11), but confirmed the absence of a spatial effect on
358 BBL seawater and sediment samples for 18S data. Analysis of homogeneity of variance between
359 APEIs found a significant difference between groups for the 18S data only ($p = 0.044$; Table S8).

360 3.3.3 | Habitat

361 The total sediment ASV and OTU gamma diversity was significantly higher (~ 2x higher) on
362 abyssal seamounts than on abyssal plains for both markers, as indicated by the absence of overlap in
363 the confidence intervals in Figure 4. When analysed per APEI, only the APEI4 seamount had a
364 significantly higher richness than the adjacent plain (Fig. S4). Community composition was
365 significantly different between habitats for both sediment ($p \leq 0.05$; Table S9) and BBL seawater
366 samples ($p \leq 0.024$; Table S9), with no significant difference in group dispersion among habitats
367 (Table S8). Relative diversity of arthropods and platyhelminths was higher in seamount sediments in
368 comparison to adjacent abyssal plains (Fig. 6), with a higher fraction of ASV diversity in nematodes
369 in abyssal plain habitats. Comparison of BBL seawater between plains and seamounts found higher
370 relative diversity of nemerteans on the plains and higher chordate diversity over seamount summits.
371 Three families occurring in at least five samples were found to be specifically associated with abyssal
372 plains: These included Nerillidae (annelid), and the nematode families Monhysteridae and
373 Comesomatidae (Table S12).

374 3.4 | Biogeography and range distributions across APEIs and habitats

375 The proportion of taxa unique to each APEI and bathymetric habitat was similar between 18S
376 and COI data (Fig. 7), and significantly higher for seamounts (mean of 90 % and 82 % for 18S and
377 COI, respectively) than abyssal plains (mean of 85 % and 72 % for 18S and COI, respectively)
378 (Kruskal-Wallis, $p < 0.001$, both markers; Table S13). The proportion of bathymetric habitat-specific
379 taxa, or those restricted to either seamounts or abyssal plains but found in different APEIs
380 (widespread-specific), was significantly lower for seamounts than abyssal plains (Kruskal-Wallis, $p <$
381 0.001 , Table S13). Conversely, taxa not specific to any habitat or APEI (cosmopolitan taxa)
382 represented a slightly larger proportion of the community at seamount summits (8.5 % and 14.5 % for
383 18S and COI, respectively) than on the abyssal plains (5.9 % and 11.2 % for 18S and COI,

384 respectively; Fig. 7). Figure 7B shows that taxa found to be widespread across APEIs but
385 bathymetrically restricted were exclusively arthropods, nematodes or unidentified metazoans.
386 Cosmopolitan taxa included these groups as well as annelids, chordates, nemerteans and flatworms.
387 Taxa unique to a habitat-APEI combination included the widest range of taxonomic groups, with
388 cnidarians, ctenophores, gastrotrichs, hemichordates and kinorhynchans in addition to the more
389 widespread groups (Fig. 7B). 26 % of COI OTUs (56 of 212) that were found to be cosmopolitan in
390 habitat association had ASVs, or COI haplotypes, that were specific to either seamount or abyssal
391 plain habitats (for COI OTUs and ASVs observed in a minimum of 5 and 3 samples, respectively; >
392 50 reads). This result suggests that approximately a quarter of cosmopolitan taxa may have population
393 genetic structure with COI haplotypes that are restricted in distribution to part of the species
394 geographic range.

395

397 Deep-sea ecosystems are under increasing anthropogenic pressure, with deep seabed mining a
398 near-term threat (Fukushima et al., 2017). Yet accurately characterizing biodiversity in the deep-sea
399 benthos using conventional surveys (e.g. visual, morpho-taxonomy) requires extensive resources
400 (Brandt et al., 2014), due to the remoteness of the habitat, challenging environmental conditions and
401 relatively high numbers of rare invertebrate taxa. In this study, we attempt to address these issues by
402 applying eDNA metabarcoding to assess metazoan diversity across substrates, habitats and large-scale
403 environmental gradients in the abyssal western Clarion Clipperton Zone.

404 Our results confirm that eDNA methods capture distinct communities as are known to occur in
405 association with different substrates in the abyssal ocean (e.g., Amon et al. 2016, DeSmet et al. 2017).
406 This observation is important because one requirement for successful application of eDNA
407 metabarcoding as a biomonitoring tool in the CCZ is that the method be sensitive enough to detect
408 distinct communities that occur in close geographic proximity. We observed very distinct
409 communities sampled in sediments, on polymetallic nodules, and in the BBL seawater (Fig. 5), with
410 little organismal overlap ($< 5\%$) among ASVs (18S) and OTUs (COI) sampled at equivalent
411 sampling coverage in distinct sample types (substrates). Taxa found exclusively on nodules were
412 mostly sessile suspension feeders, including bryozoans, alcyonacean corals (Isididae), ascidians
413 (Styelidae), brachiopods (Terebratulida), a number of sponge taxa (within Cladorhizidae,
414 Hexactinellida, and Suberitida), and bivalves (Venerida, Mytilida), among others (Tables S5, S6), and
415 this organismal list is broadly similar to nodule-attached metazoans reported in prior work
416 (Mullineaux et al. 1987, Veillette et al. 2007, Amon et al. 2016, Vanreusel et al. 2016). Taxa
417 simultaneously found in association with both sediments and nodules were predominantly mobile
418 organisms, including nematodes, arthropods and annelids, with the exception of a few sessile families,
419 such as Arcidae (bivalve), Cladorhizidae (sponge), and Hexacrobrylidae (ascidian). Organisms
420 sampled exclusively in sediments were overwhelmingly nematodes (79 18S ASVs of 197 total ASVs
421 that were exclusive to sediments), the dominant meiofaunal phylum. Although we expected that BBL
422 plankton eDNA might settle to the seafloor, very few BBL ASVs and OTUs were observed in
423 sediments ($< 6\%$) or nodules ($\leq 2\%$).

424 We find evidence that abyssal seamounts may represent biodiversity hotspots for benthic
425 organisms (e.g., 1.4 to 2.4 times higher richness, APEI 4), with distinct community composition and
426 community biogeography in comparison to the adjacent abyssal plains in the western CCZ.
427 Seamounts have long been hypothesized to be species richness hotspots (e.g., McClain 2007), but
428 evidence to support this hypothesis has been mixed (Rowden et al. 2010a), with several studies
429 finding equivalent or lower richness on seamounts than on slopes or adjacent non-seamount areas
430 (e.g., fishes, megafauna; Tracey et al. 2004, O'Hara et al. 2007, Howell et al. 2010). Results from this
431 study provide new insights into the potential role of seamounts as biodiversity hotspots in that (1) our
432 observations derive from seamounts that are more remote and with abyssal summit depths (~3100,
433 3500 m) that are deeper than the vast majority of seamounts studied to date, and (2) we use eDNA
434 metabarcoding to estimate ASV/OTU richness, yielding greater taxonomic coverage and greater
435 emphasis on smaller, more cryptic organisms than studies using conventional survey techniques. Our
436 genetic eDNA data also have the asset that our observations are not limited by the current state of
437 taxonomic knowledge for the assemblage. Given that > 80 % of macrofaunal and meiofaunal
438 invertebrates at abyssal depths are undescribed (Snelgrove & Smith 2002, George et al. 2014), this is
439 a considerable strength over morphology-based measures. A number of mechanisms could cause
440 elevated richness on seamounts, including higher habitat heterogeneity and/or heightened beta
441 diversity reflecting faunal turnover across depth along the seamount flank, increased trophic input that
442 supports elevated invertebrate abundance, biomass, and diversity, or increased speciation rates due to
443 the geographic isolation of seamounts (among others; McLain 2007, Rowden, Clark et al., 2010,
444 Zeppilli et al. 2014). The few prior quantitative studies of meiofaunal assemblages on seamounts have
445 found that although summits may not have elevated richness relative to flanks or adjacent abyssal
446 plain areas, they do have a very distinct nematode/copepod assemblage, with many species that are
447 bathymetrically restricted in range and with high faunal turnover across depth and substrate on the
448 seamount flank (enhancing beta diversity; George 2013, Zeppilli et al. 2013, Zeppilli et al. 2014,
449 George et al. 2018). Our results regarding distinct sediment community composition on seamount
450 summits (Fig. 6), largely driven by meiofaunal taxa, are broadly congruent with these prior
451 observations. In the case of eDNA, one additional possible mechanism driving higher richness on
452 seamounts is that seamount eDNA samples may integrate a larger spatial area than those on the

453 plains, with bedload transport importing particulate matter and eDNA from microhabitat patches
454 elsewhere on the seamount (beta diversity). Seamount summits are physically more open systems than
455 abyssal plains, often with higher turbulence and current velocities (White et al., 2008), and eDNA
456 may be transported into a site from nearby habitat patches. In this study, inference of the true richness
457 on seamounts was constrained by the limited sampling coverage achieved (< 30 %). Further research
458 is needed to confirm the hypothesis that seamounts are biodiversity hotspots across the abyss.

459 Seamounts have historically been perceived as isolated habitats, possibly harbouring high
460 levels of endemism, due to their geographic isolation and hydrographic peculiarities (e.g. Taylor
461 column formation), which can hinder larval dispersal and limit connectivity among populations (Clark
462 et al., 2010; McClain et al., 2009; Samadi et al., 2006). Although limited evidence has been found
463 supporting the seamount endemism hypothesis (McClain et al. 2009; Rowden et al., 2010b), our
464 results suggest that abyssal seamount benthic communities display less connectivity between APEIs
465 than comparable communities on the abyssal plain. Specifically, a smaller proportion of the seamount
466 community is comprised of taxa that are bathymetrically-restricted but widespread across APEIs
467 (seamount-associated) than is observed for abyssal plain assemblages (plains-associated). In other
468 words, most seamount taxa with broad biogeographic ranges were not specific to a particular
469 bathymetric habitat (seamounts, plains). In direct contrast, the majority of widespread (observed
470 across different APEIs) abyssal plain taxa were not observed on seamounts and therefore may lack the
471 capacity to colonize them. We also observe that a higher fraction of the seamount fauna is unique to
472 habitat and APEI (endemics & pseudo-endemics) than in abyssal plain habitats, at equivalent
473 sampling coverage. In addition, several cosmopolitan OTUs were composed of sequence variants, or
474 COI haplotypes, that were associated with a specific bathymetric habitat; this is initial tentative
475 evidence of population genetic differentiation between plain and seamount populations within these
476 putative species (26 % of cosmopolitan taxa). Collectively, these observations support the inference
477 that seamounts likely act both as biogeographic islands for taxa with limited dispersal ability, but also
478 as stepping stones for dispersal for more cosmopolitan taxa (Miller & Gunasekera, 2017; Rowden et
479 al., 2010b). Other studies report mixed support for seamounts as stepping stones for dispersal (e.g.
480 Wilson & Kaufman 1987, O'Hara et al. 2010), and taxon-specific traits related to dispersal ability
481 likely drive these broader biogeographic trends.

482 Abyssal ecosystems are strongly modulated by the flux of detrital material originating from
483 the upper ocean due to food limitation in the abyss (Smith et al. 2008). Both abundance and diversity
484 of macrofaunal invertebrates have been shown to positively correlate with POC flux (De Smet et al.,
485 2017; Rex et al., 2006; Smith et al., 1997). Polymetallic nodules also enhance the abundance and
486 regional diversity of the deep-sea benthos as they provide hard substrate in an otherwise soft-bottom
487 environment for a range of sessile epifauna (Veillette et al. 2007, Amon et al., 2016; Vanreusel et al.,
488 2016). APEIs sampled in this study span a range of moderate to low POC flux (Tables 2, S2; Lutz et
489 al. 2007; Wedding et al., 2013; Smith et al., 2019) and high to low polymetallic nodule abundance
490 (Table 2; Morgan et al. 2010; Smith et al. 2019, McQuaid et al *in review*). Overall, taxon richness was
491 lowest within APEI 1, at lowest POC flux, and highest within APEI 4, at moderate POC flux and in a
492 region containing both soft sediment habitat and high nodule abundance. Significant differences in
493 sediment-community composition were observed between APEIs. While spatial distance may be
494 partly responsible for these differences, at least in the COI data, these results support the idea that
495 POC flux and/or nodule density positively affect community diversity. We also find that nodule size,
496 measured here as weight, influenced community composition. While this relationship was not
497 observed in De Smet et al. (2017), it is concordant with results from Simon-Lledó et al. (2019),
498 suggesting nodule-size preferences among taxa.

499 eDNA metabarcoding could be a powerful and cost-effective method of assessing biodiversity
500 in baseline surveys of the deep-sea. However, one of the primary limitations is the low representation
501 of deep-sea organisms in reference sequence databases (Lacoursière-Roussel et al., 2018;
502 Wangenstein, Palacín, Guardiola, & Turon, 2018; Kersten et al. 2019). In this study, only 25 % and
503 1.5 % of 18S and COI metazoan sequences could be assigned to Family. This problem was especially
504 pronounced in the COI data, where ~ 19 % of metazoan reads could only be assigned to phylum.
505 While many of the unassigned sequences likely derive from undescribed organisms that are new to
506 science, a large fraction likely also corresponds to fully described taxa that lack representative DNA
507 barcodes (see Lacoursière-Roussel et al., 2018). The absence of taxonomic, and therefore ecological
508 information, hinders our capacity to understand deep-sea ecosystem processes and design and
509 implement effective conservation measures. It is imperative that we continue allocating time and
510 resources to describing new species, and augmenting reference databases with DNA barcodes for

511 described species (Glover et al., 2018). Given our results, efforts should be directed towards the
512 characterization of meiofaunal taxa in particular, as there is very high, but unclassified, diversity in
513 sediments.

514 5 | Conclusions and Management Implications

515 Our results suggest that abyssal seamounts are important reservoirs of metazoan diversity in
516 the abyssal CCZ, with elevated taxon richness relative to abyssal plains habitats. We observed distinct
517 community composition on seamounts (as in Zeppilli et al. 2013, 2014 and George et al. 2018), and
518 limited taxonomic overlap with the adjacent abyssal plain assemblages (499 of OTUs [16 %] and 379
519 of OTUs [19 %] for APEIs 4 and 7, respectively), implying that even if seamount populations persist
520 within claim areas during large-scale seabed mining, they will not serve as major source populations
521 to reseed disturbed areas of the adjacent abyssal plains. Conservation of these biologically distinct
522 communities is important, but insufficient to ensure preservation of viable populations of the
523 dominant abyssal plain fauna. We observed fairly large range distributions (up to 1500 km) for 2.4 %
524 of the plains fauna (COI OTUs cosmopolitan across APEIs 1, 4, 7 and present in at least 5 samples),
525 suggesting that some species are distributed across spatial scales bridging APEIs and claim areas.
526 The majority of OTUs/ASVs, however, were rare and limited to small spatial areas in our material,
527 and so we cannot reject the hypothesis that they have restricted species ranges. In accordance with
528 other studies, we also find highest metazoan richness in regions with both substantial nodule cover
529 and soft sediment habitats, as well as moderate POC flux, environmental variables that have been
530 shown to correlate with higher abundance and diversity of megafaunal invertebrates within the CCZ
531 (e.g. Amon et al., 2016; De Smet et al., 2017; Vanreusel et al., 2016). Finally, in this first eDNA study
532 for the western Clarion Clipperton Zone, we demonstrate that eDNA metabarcoding could be a
533 powerful survey tool for assessing community diversity in the context of seabed mining impacts. The
534 taxonomic resolution is comparable to or higher than that typically obtained using image-based
535 survey techniques, and the communities detected are tightly linked to substrate type (nodules,
536 sediments). Additional efforts to expand reference databases through DNA barcoding will enhance
537 the classification power of eDNA methods, enabling more useful assessments and testing of long-
538 standing deep-sea ecological hypotheses.

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549 7 | Author contributions

550 EG, OK, OL, and CS designed the study. EG, OK, and CS conducted field work and sampling
551 at sea. OL generated the data and performed analyses, and OL wrote the manuscript with intellectual
552 contributions from all co-authors. EG and CS provided grant and equipment support.

553 8 | Data Accessibility

554 Unprocessed sequences are accessible from the NCBI Sequence Read Archive (SRA) under
555 accession numbers SRR9199590 to SRR9199853. Metadata for the samples are available in the
556 supplementary information.

557

References

- Amon, D. J., Ziegler, A. F., Dahlgren, T. G., Glover, A. G., Goineau, A., Gooday, A. J., ... Smith, C. R. (2016). Insights into the abundance and diversity of abyssal megafauna in a polymetallic-nodule region in the eastern Clarion-Clipperton Zone. *Scientific Reports*, 6(1), 30492. doi: 10.1038/srep30492
- Asnicar, F., Weingart, G., Tickle, T. L., Huttenhower, C., & Segata, N. (2015). Compact graphical representation of phylogenetic data and metadata with GraPhlAn. *PeerJ*, 3, e1029. doi: 10.7717/peerj.1029
- Beckmann, A., & Mohn, C. (2002). The upper ocean circulation at Great Meteor Seamount. *Ocean Dynamics*, 52(4), 194–204. doi: 10.1007/s10236-002-0018-3
- Benson, D. A., Karsch-mizrachi, I., Lipman, D. J., Ostell, J., & Wheeler, D. L. (2008). *GenBank*. 36(December 2007), 25–30. doi: 10.1093/nar/gkm929
- Bokulich, N. A., Kaehler, B. D., Rideout, J. R., Dillon, M., Bolyen, E., Knight, R., ... Gregory Caporaso, J. (2018). Optimizing taxonomic classification of marker-gene amplicon sequences with QIIME 2's q2-feature-classifier plugin. *Microbiome*, 6(1), 1–17. doi: 10.1186/s40168-018-0470-z
- Boschen, R. E., Collins, P. C., Tunnicliffe, V., Carlsson, J., Gardner, J. P. A., Lowe, J., ... Swaddling, A. (2016). A primer for use of genetic tools in selecting and testing the suitability of set-aside sites protected from deep-sea seafloor massive sulfide mining activities. *Ocean and Coastal Management*, 122, 37–48. doi: 10.1016/j.ocecoaman.2016.01.007
- Boylen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C., Ghalith, G. A. Al, ... Naimey, A. T. (2018). QIIME 2 : Reproducible , interactive , scalable , and extensible microbiome data science. *PeerJ Preprints*. doi: 10.7287/peerj.preprints.27295v1
- Brandt, A., Grif Ths, H., Gutt, J., Linse, K., Schiaparelli, S., Ballerini, T., ... Pfannkuche, O. (2014). Challenges of deep-sea biodiversity assessments in the Southern Ocean. *Adv Polar Sci*, 25(3), 204–212. doi: 10.13679/j.advps.2014.3.00204
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes, S. P. (2016).

DADA2: High-resolution sample inference from Illumina amplicon data. *Nature Methods*, 13(7), 581–583. doi: 10.1038/nmeth.3869

Camacho, C., Coulouris, G., Avagyan, V., Ma, N., Papadopoulos, J., Bealer, K., & Madden, T. L. (2009). BLAST+: architecture and applications. *BMC Bioinformatics*, 10(1), 421. doi: 10.1186/1471-2105-10-421

Chao, A., Gotelli, N. J., Hsieh, T. C., Sander, E. L., Ma, K. H., Colwell, R. K., & Ellison, A. M. (2014). Rarefaction and extrapolation with Hill numbers: A framework for sampling and estimation in species diversity studies. *Ecological Monographs*, 84(1), 45–67. doi: 10.1890/13-0133.1

Cho, W., & Shank, T. M. (2010). Incongruent patterns of genetic connectivity among four ophiuroid species with differing coral host specificity on North Atlantic seamounts. *Marine Ecology*, 31(SUPPL. 1), 121–143. doi: 10.1111/j.1439-0485.2010.00395.x

Clark, M. R., Rowden, A. A., Schlacher, T., Williams, A., Consalvey, M., Stocks, K. I., ... Hall-Spencer, J. M. (2010). The Ecology of Seamounts: Structure, Function, and Human Impacts. *Annual Review of Marine Science*, 2(1), 253–278. doi: 10.1146/annurev-marine-120308-081109

Cordier, T., Esling, P., Lejzerowicz, F., Visco, J., Ouadahi, A., Martins, C., ... Pawlowski, J. (2017). Predicting the ecological quality status of marine environments from eDNA metabarcoding data using supervised machine learning. *Environ. Sci. Technol.*, 51(16), 9118-9126. doi: 10.1021/acs.est.7b01518

Cristescu, M. E., & Hebert, P. D. N. (2018). Uses and Misuses of Environmental DNA in Biodiversity Science and Conservation. *Annual Review of Ecology, Evolution, and Systematics*, 49(1), 209–230. doi: 10.1146/annurev-ecolsys-110617-062306

Danovaro, R., Carugati, L., Berzano, M., Cahill, A. E., Carvalho, S., Chenuil, A., ... Borja, A. (2016). Implementing and Innovating Marine Monitoring Approaches for Assessing Marine Environmental Status. *Frontiers in Marine Science*, 3(November), 213. doi: 10.3389/fmars.2016.00213

Dawson, M. N. (2016). Island and island-like marine environments. *Global Ecology and Biogeography*, 25(7), 831–846. doi: 10.1111/geb.12314

De Smet, B., Pape, E., Riehl, T., Bonifácio, P., Colson, L., & Vanreusel, A. (2017). The Community

Structure of Deep-Sea Macrofauna Associated with Polymetallic Nodules in the Eastern Part of the Clarion-Clipperton Fracture Zone. *Frontiers in Marine Science*, 4(April), 1–14. doi: 10.3389/fmars.2017.00103

Dell'Anno, A., Carugati, L., Corinaldesi, C., Riccioni, G., & Danovaro, R. (2015). Unveiling the Biodiversity of Deep-Sea Nematodes through Metabarcoding: Are We Ready to Bypass the Classical Taxonomy? *PLoS ONE*, 10(12), e0144928. doi: 10.1371/journal.pone.0144928

Everett, M. V., & Park, L. K. (2018). Exploring deep-water coral communities using environmental DNA. *Deep Sea Research Part II: Topical Studies in Oceanography*, 150, 229–241. doi: 10.1016/j.dsr2.2017.09.008

Foster, Z., Sharpton, T., & Grünwald, N. (2017). Metacoder: An R package for visualization and manipulation of community taxonomic diversity data. *PLOS Computational Biology*, 13(2), 1–15. doi: 10.1371/journal.pcbi.1005404

Fukushima, T., & Nishijima, M. (2017). Taxonomic Problems in Environmental Impact Assessment (EIA) Linked to Ocean Mining and Possibility of New Technology Developments. In R. Sharma (Ed.), *Deep-Sea Mining* (pp. 465–482). doi: 10.1007/978-3-319-52557-0_16

Geller, J., Meyer, C., Parker, M., & Hawk, H. (2013). Redesign of PCR primers for mitochondrial cytochrome c oxidase subunit I for marine invertebrates and application in all-taxa biotic surveys. *Molecular Ecology Resources*, 13(5), 851–861. doi: 10.1111/1755-0998.12138

Genin, A., & Dower, J. F. (2007). Seamounts: Ecology, Fisheries & Conservation. In T. J. Pitcher, T. Morato, P. J. B. Hart, M. R. Clark, N. Haggan, & R. S. Santos (Eds.), *Seamounts: Ecology, Fisheries & Conservation*. doi: 10.1002/9780470691953

George, K. H., Veit-Köhler, G., Arbizu, P. M., Seifried, S., Rose, A., Willen, E., ... Schminke, H. K. (2014). Community structure and species diversity of Harpacticoida (Crustacea: Copepoda) at two sites in the deep sea of the Angola Basin (Southeast Atlantic). *Organisms Diversity & Evolution*, 14(1), 57–73. doi: 10.1007/s13127-013-0154-2

Glover, A. G., & Smith, C. R. (2003). The deep-sea floor ecosystem: current status and prospects of anthropogenic change by the year 2025. *Environmental Conservation*, 30(3), 219–241. doi: 10.1017/S0376892903000225

Glover, A. G., Wiklund, H., Chen, C., & Dahlgren, T. G. (2018). Managing a sustainable deep-sea

'blue economy' requires knowledge of what actually lives there. *ELife*, 7, 1–7. doi: 10.7554/eLife.41319

George, K. H. (2013). Faunistic research on metazoan meiofauna from seamounts - a review. *Meiofauna Marina*, 20 (February), 1–32.

George, K. H., Pointner, K., & Packmor, J. (2018). The benthic Copepoda (Crustacea) of Anaximenes Seamount (eastern Mediterranean Sea)—Community structure and species distribution. *Progress in Oceanography*, 165, 299–316. <https://doi.org/10.1016/j.pocean.2018.06.006>

Gollner, S., Kaiser, S., Menzel, L., Jones, D. O. B., Brown, A., Mestre, N. C., ... Martinez Arbizu, P. (2017). Resilience of benthic deep-sea fauna to mining activities. *Marine Environmental Research*, 129, 76–101. doi: 10.1016/j.marenvres.2017.04.010

Goodwin, K. D., Thompson, L. R., Duarte, B., Kahlke, T., Thompson, A. R., Marques, J. C., & Caçador, I. (2017). DNA Sequencing as a Tool to Monitor Marine Ecological Status. *Frontiers in Marine Science*, 4(107). doi: 10.3389/fmars.2017.00107

Guardiola, M., Uriz, M. J., Taberlet, P., Coissac, E., Wangensteen, O. S., & Turon, X. (2015). Deep-Sea, Deep-Sequencing: Metabarcoding Extracellular DNA from Sediments of Marine Canyons. *PLoS ONE*, 10(10), e0139633. doi: 10.1371/journal.pone.0139633

Guardiola, M., Wangensteen, O. S., Taberlet, P., Coissac, E., Uriz, M. J., & Turon, X. (2016). Spatio-temporal monitoring of deep-sea communities using metabarcoding of sediment DNA and RNA. *PeerJ*, 4(December), e2807. doi: 10.7717/peerj.2807

Hall-Spencer, J. M., Rogers, A. D., Davies, J., & Foggo, A. (2007). Deep-sea coral distribution on seamounts, oceanic islands, and continental slopes in the Northeast Atlantic. *Bulletin of Marine Science*, 81(3), 135–146.

Hannides, A. K., & Smith, C. R. (2003). The Northeastern Pacific Abyssal Plain. In K. D. Black & G. B. Shimmield (Eds.), *Biogeochemistry of Marine Systems* (pp. 208–237). Blackwell Publishing Ltd.

Harris, P. T., Macmillan-Lawler, M., Rupp, J., & Baker, E. K. (2014). Geomorphology of the oceans. *Marine Geology*, 352, 4–24. doi: 10.1016/j.margeo.2014.01.011

Howell, K. L., Mowles, S. L., & Foggo, A. (2010). Mounting evidence: near-slope seamounts are faunally indistinct from an adjacent bank. *Marine Ecology*, 31(SUPPL. 1), 52–62.

<https://doi.org/10.1111/j.1439-0485.2010.00368.x>

Hsieh, T. C., Ma, K. H., & Chao, A. (2016). iNEXT: an R package for rarefaction and extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution*, 7(12), 1451–1456. doi: 10.1111/2041-210X.12613

Huvenne, V. A. I., Bett, B. J., Masson, D. G., Le Bas, T. P., & Wheeler, A. J. (2016). Effectiveness of a deep-sea cold-water coral Marine Protected Area, following eight years of fisheries closure. *Biological Conservation*, 200(June), 60–69. doi: 10.1016/j.biocon.2016.05.030

ISA. (2019). *Deep CCZ Biodiversity Synthesis Workshop*. Friday Harbor, Washington, USA: International Seabed Authority.

Jones, D. O. B., Amon, D. J., & Chapman, A. S. A. (2018). Mining Deep-Ocean Mineral Deposits: What are the Ecological Risks? *Elements*, 14(5), 325–330. <https://doi.org/10.2138/gselements.14.5.325>

Kaiser, S., Smith, C. R., & Arbizu, P. M. (2017). Editorial: Biodiversity of the Clarion Clipperton Fracture Zone. *Marine Biodiversity*, 47(2), 259–264. doi: 10.1007/s12526-017-0733-0

Kassambara, A. (2018). *ggpubr: “ggplot2” Based Publication Ready Plots*. Retrieved from <https://cran.r-project.org/package=ggpubr>

Katoh, K., & Standley, D. M. (2013). *MAFFT Multiple Sequence Alignment Software Version 7: Improvements in Performance and Usability Article Fast Track*. 30(4), 772–780. doi: 10.1093/molbev/mst010

Kersten, O., Vetter, E. W., Jungbluth, M. J., Smith, C. R., & Goetze, E. (2019). Larval assemblages over the abyssal plain in the Pacific are highly diverse and spatially patchy. *PeerJ*, 7, e7691. <https://doi.org/10.7717/peerj.7691>

Koslow, J. A., Gowlett-Holmes, K., Lowry, J. K., O’Hara, T., Poore, G. C. B., & Williams, A. (2001). Seamount benthic macrofauna off southern Tasmania: Community structure and impacts of trawling. *Marine Ecology Progress Series*, 213(April), 111–125. doi: 10.3354/meps213111

Lacoursière-Roussel, A., Howland, K., Normandeau, E., Grey, E. K., Archambault, P., Deiner, K., ... Bernatchez, L. (2018). eDNA metabarcoding as a new surveillance approach for coastal Arctic biodiversity. *Ecology and Evolution*, 8(16), 7763–7777. doi: 10.1002/ece3.4213

Laroche, O., Kersten, O., Smith, C. R., Goetze, E. (2020). From sea surface to seafloor: a benthic

- allochthonous eDNA survey for the abyssal ocean. *bioRxiv*. doi: 10.1101/2020.05.07.082602.
- Larsson, J. (2019). *{eulerr}: Area-Proportional {Euler} and {Venn} Diagrams with Ellipses*. Retrieved from <https://cran.r-project.org/package=eulerr>
- Leal, J. H., & Bouchet, P. (1991). Distribution patterns and dispersal of prosobranch gastropods along a seamount chain in the Atlantic Ocean. *Journal of the Marine Biological Association of the United Kingdom*, 71(1), 11–25. doi: 10.1017/S0025315400037358
- Lejzerowicz, F., Esling, P., & Pawlowski, J. (2014). Patchiness of deep-sea benthic Foraminifera across the southern ocean: Insights from High-throughput DNA sequencing. *Deep Sea Research Part II: Topical Studies in Oceanography*, 108, 17–26. doi: 10.1016/j.dsr2.2014.07.018
- Leray, M., Yang, J. Y., Meyer, C. P., Mills, S. C., Agudelo, N., Ranwez, V., ... Machida, R. J. (2013). A new versatile primer set targeting a short fragment of the mitochondrial COI region for metabarcoding metazoan diversity: application for characterizing coral reef fish gut contents. *Frontiers in Zoology*, 10(1), 34. doi: 10.1186/1742-9994-10-34
- Lozupone, C., & Knight, R. (2005). UniFrac : a New Phylogenetic Method for Comparing Microbial Communities. *Applied and Environmental Microbiology*, 71(12), 8228–8235. doi: 10.1128/AEM.71.12.8228
- Lutz, M. J., Caldeira, K., Dunbar, R. B., & Behrenfeld, M. J. (2007). Seasonal rhythms of net primary production and particulate organic carbon flux to depth describe the efficiency of biological pump in the global ocean. *Journal of Geophysical Research*, 112(C10), C10011. doi: 10.1029/2006JC003706
- Machida, R. J., Leray, M., Ho, S.-L., & Knowlton, N. (2017). Metazoan mitochondrial gene sequence reference datasets for taxonomic assignment of environmental samples. *Scientific Data*, 4(March), 1–7. doi: 10.1038/sdata.2017.27
- Martin, M. (2011). Cutadapt removes adapter sequences from high-throughput sequencing reads. *EMBnet.Journal*, 17(1), 10. doi: 10.14806/ej.17.1.200
- McClain, C. R. (2007). Seamounts: identity crisis or split personality? *Journal of Biogeography*, 34(12), 2001–2008. doi: 10.1111/j.1365-2699.2007.01783.x
- McClain, C. R., Lundsten, L., Ream, M., Barry, J., & DeVogelaere, A. (2009). Endemicity, Biogeography, Composition, and Community Structure On a Northeast Pacific Seamount. *PLoS*

ONE, 4(1), e4141. doi: 10.1371/journal.pone.0004141

McMurdie, P. J., & Holmes, S. (2013). phyloseq: an R package for reproducible interactive analysis and graphics of microbiome census data. *PLoS ONE*, 8(4), e61217. doi:

10.1371/journal.pone.0061217

McQuaid, K. A., Attrill, M., Cobley, A., Glover, A. G., Smith, C. R., Howell, K. L. (in review). Using a top-down, broad-scale habitat classification to assess representivity of the CCZ APEI network.

Miller, K. A., Thompson, K. F., Johnston, P., & Santillo, D. (2019). An Overview of Seabed Mining Including the Current State of Development , Environmental Impacts , and Knowledge Gaps.

Front. Mar. Sci., 4(January 2018). doi: 10.3389/fmars.2017.00418

Miller, K. J., & Gunasekera, R. M. (2017). A comparison of genetic connectivity in two deep sea corals to examine whether seamounts are isolated islands or stepping stones for dispersal.

Scientific Reports, 7(April), 1–14. doi: 10.1038/srep46103

Morato, T., Hoyle, S. D., Allain, V., & Nicol, S. J. (2010). Seamounts are hotspots of pelagic biodiversity in the open ocean. *Proceedings of the National Academy of Sciences*, 107(21),

9707–9711. doi: 10.1073/pnas.0910290107

Morgan, C., Kotlinski, R., Stoyanova, V., Zhou, H., Lu, W., Zhou, N., Li, D., Yubko, V., Parson, L., Hunter, P. M., Smith, C. R., Mincks, S., Kang, J-K., Hyeong, K., Park, C-K., Tak Ko, Y., Kim, J., Yang, S., Cronan, D. S., Kazmin, Y., Hoffert, M. (2010) A geological model of polymetallic nodule deposits in the Clarion Clipperton Fracture Zone. *International Seabed Authority (ISA), Technical Study: No. 6*, Kingston, Jamaica. Workshop 2006

Mullineaux, L. S. (1987). Organisms living on manganese nodules and crusts: distribution and abundance at three North Pacific sites. *Deep Sea Research Part A, Oceanographic Research Papers*, 34(2), 165–184. doi: 10.1016/0198-0149(87)90080-X

O'Hara, T. D., Consalvey, M., Lavrado, H. P., & Stocks, K. I. (2010). Environmental predictors and turnover of biota along a seamount chain. *Marine Ecology*, 31(SUPPL. 1), 84–94. doi:

10.1111/j.1439-0485.2010.00379.x

Oebius, H. U., Becker, H. J., Rolinski, S., & Jankowski, J. A. (2001). Parametrization and evaluation of marine environmental impacts produced by deep-sea manganese nodule mining. *Deep Sea Research Part II: Topical Studies in Oceanography*, 48(17–18), 3453–3467. doi:

10.1016/S0967-0645(01)00052-2

Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., ... Wagner, H. (2018). *vegan: Community Ecology Package*. Retrieved from <https://cran.r-project.org/package=vegan>

Pagès, H., Aboyoun, P., Gentleman, R., & DebRoy, S. (2019). *Biostrings: Efficient manipulation of biological strings*. R package version 2.52.0. doi: 10.18129/B9.bioc.Biostrings

Price, M. N., Dehal, P. S., & Arkin, A. P. (2010). *FastTree 2 – Approximately Maximum-Likelihood Trees for Large Alignments*. 5(3). doi: 10.1371/journal.pone.0009490

Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., ... Glöckner, F. O. (2013). The SILVA ribosomal RNA gene database project: Improved data processing and web-based tools. *Nucleic Acids Research*, 41(D1), 590–596. doi: 10.1093/nar/gks1219

Radziejewska, T. (2014). Meiobenthos in the Sub-equatorial North-Eastern Pacific Abyssal Seafloor. In *Meiobenthos in the Sub-equatorial Pacific Abyss* (pp.29-65). Springer, Berlin, Heidelberg. doi: 10.1007/978-3-642-41458-9

Ramirez-Llodra, E., Brandt, A., Danovaro, R., De Mol, B., Escobar, E., German, C. R., ... Vecchione, M. (2010). Deep, diverse and definitely different: unique attributes of the world's largest ecosystem. *Biogeosciences*, 7(9), 2851–2899. doi: 10.5194/bg-7-2851-2010

Rex, M., Etter, R., Morris, J., Crouse, J., McClain, C., Johnson, N., ... Avery, R. (2006). Global bathymetric patterns of standing stock and body size in the deep-sea benthos. *Marine Ecology Progress Series*, 317(January), 1–8. doi: 10.3354/meps317001

Rogers, A. D. (1994). The Biology of Seamounts. In *Advances in marine biology* (Vol. 30, pp. 305–350). doi: 10.1016/S0065-2881(08)60065-6

Rognes, T., Flouri, T., Nichols, B., Quince, C., & Mahé, F. (2016). VSEARCH: a versatile open source tool for metagenomics. *PeerJ*, 4, e2584. doi: 10.7717/peerj.2584

Rowden, A. A., Schlacher, T. A., Williams, A., Clark, M. R., Stewart, R., Althaus, F., ... Dowdney, J. (2010a). A test of the seamount oasis hypothesis: Seamounts support higher epibenthic megafaunal biomass than adjacent slopes. *Marine Ecology*, 31(SUPPL. 1), 95–106. doi: 10.1111/j.1439-0485.2010.00369.x

Rowden, A. A., Dower, J. F., Schlacher, T. A., Consalvey, M., & Clark, M. R. (2010b). Paradigms in

seamount ecology: Fact, fiction and future. *Marine Ecology*, 31(SUPPL. 1), 226–241. doi: 10.1111/j.1439-0485.2010.00400.x

Samadi, S., Botton, L., Macpherson, E., De Forges, B. R., Boisselier, M.-C., Botton, L., & Macpherson, E. (2006). Seamount endemism questioned by the geographic distribution and population genetic structure of marine invertebrates. *Marine Biology*, 149(6), 1463–1475. doi: 10.1007/s00227-006-0306-4

Shulse, C. N., Maillot, B., Smith, C. R., & Church, M. J. (2017). Polymetallic nodules, sediments, and deep waters in the equatorial North Pacific exhibit highly diverse and distinct bacterial, archaeal, and microeukaryotic communities. *MicrobiologyOpen*, 6(2), e00428. <https://doi.org/10.1002/mbo3.428>

Simon-Lledó, E., Bett, B. J., Huvenne, V. A. I., Schoening, T., Benoist, N. M. A., & Jones, D. O. B. (2019). Ecology of a polymetallic nodule occurrence gradient: Implications for deep-sea mining. *Limnology and Oceanography*, 1–12. doi: 10.1002/lno.11157

Sinniger, F., Pawlowski, J., Harii, S., Gooday, A. J., Yamamoto, H., Chevalloné, P., ... Creer, S. (2016). Worldwide Analysis of Sedimentary DNA Reveals Major Gaps in Taxonomic Knowledge of Deep-Sea Benthos. *Frontiers in Marine Science*, 3(June), 92. doi: 10.3389/fmars.2016.00092

Smith, C. R., Berelson, W., Demaster, D. J., Dobbs, F. C., Hammond, D., Hoover, D. J., ... Stephens, M. (1997). Latitudinal variations in benthic processes in the abyssal equatorial Pacific: control by biogenic particle flux. *Deep Sea Research Part II: Topical Studies in Oceanography*, 44(9–10), 2295–2317. doi: 10.1016/S0967-0645(97)00022-2

Smith, De Leo, F., Bernardino, A., Sweetman, A., & Martinez Arbizu, P. (2008). Abyssal food limitation, ecosystem structure and climate change. *Trends in Ecology & Evolution*, 23(9), 518–528. doi: 10.1016/j.tree.2008.05.002

Smith, C. R., Clark, M., Amon, D., Bonifácio, P., Bribiesca-Contreras, G., Christodoulou, M., Church, M., Cuvlier, D., Dahlgren, T., Drazen, J. C., Durden, J. M., Fukushima, T., Glover, A., Goetze, E., Gooday, A. J., Howell, K., Hwan, Y. O., Jones, D. O. B., Laming, S., Leitner, A., Lejzerowicz, F., Lim, S. C., Martinez Arbizu, P., McQuaid, K., Menot, L., Orcutt, B., Simon-Lledó, E., Pape, E., Ju, S.-J., Stedman, G., Stratmann, T., Sweetman, A., Vanreusel, A.,

- Accepted Article
- Washburn, T., Wear, E. K., Wenzhoefer, F., Yeeting, K., Young, R., Zeppilli, D. (2019). Deep CCZ Biodiversity Synthesis Workshop Report. International Seabed Authority (ISA). Friday Harbor, Washington, Oct 2019. https://ran-s3.s3.amazonaws.com/isa.org.jm/s3fs-public/files/documents/deep_ccz_biodiversity_synthesis_workshop_report_-_final.pdf
- Snelgrove, P., & Smith, C. (2002). A Riot of Species in An Environmental Calm. In R. N. Gibson, M. Barnes, & R. J. A. Atkinson (Eds.), *Oceanography and marine biology: an annual review* (pp. 311–342). doi: 10.1201/9780203180594.ch6
- Stat, M., Huggett, M. J., Bernasconi, R., DiBattista, J. D., Berry, T. E., Newman, S. J., ... Bunce, M. (2017). Ecosystem biomonitoring with eDNA: metabarcoding across the tree of life in a tropical marine environment. *Scientific Reports*, 7(1), 12240. doi: 10.1038/s41598-017-12501-5
- Stocks, K. I., & Hart, P. J. B. (2007). Biogeography and biodiversity of seamounts. In *Seamounts: ecology, fisheries, and conservation. Blackwell Fisheries and Aquatic Resources Series Vol. 12*, pp. 255–28. Wiley Online Library.
- Thiel, H., Schriever, G., Bussau, C., & Borowski, C. (1993). Manganese nodule crevice fauna. *Deep Sea Research Part I: Oceanographic Research Papers*, 40(2), 419–423. doi: 10.1016/0967-0637(93)90012-R
- Thiel, H., Schriever, G., Ahnert, A., Bluhm, H., Borowski, C., & Vopel, K. (2001). The large-scale environmental impact experiment DISCOL—reflection and foresight. *Deep Sea Research Part II: Topical Studies in Oceanography*, 48(17–18), 3869–3882. doi: 10.1016/S0967-0645(01)00071-6
- Tilot, V., Ormond, R., Moreno Navas, J., & Catalá, T. S. (2018). The Benthic Megafaunal Assemblages of the CCZ (Eastern Pacific) and an Approach to their Management in the Face of Threatened Anthropogenic Impacts. *Frontiers in Marine Science*, 5(February), 1–25. doi: 10.3389/fmars.2018.00007
- Tracey, D. M., Bull, B., Clark, M. R., & MaCkay, K. A. (2004). Fish species composition on seamounts and adjacent slope in New Zealand waters. *New Zealand Journal of Marine and Freshwater Research*, 38(1), 163–182. <https://doi.org/10.1080/00288330.2004.9517226>
- Vanreusel, A., Hilario, A., Ribeiro, P. A., Menot, L., & Arbizu, P. M. (2016). Threatened by mining, polymetallic nodules are required to preserve abyssal epifauna. *Scientific Reports*, 6(1), 26808.

doi: 10.1038/srep26808

- Veillette, J., Sarrazin, J., Gooday J., A., Galeron, J., Caprais, J.-C., Vangriesheim, A., ... Juniper S., K. (2007). Ferromanganese nodule fauna in the Tropical North Pacific Ocean: Species richness, faunal cover and spatial distribution. *Deep-Sea Research*, 54, 1912–1935. doi: 10.1016/j.dsr.2007.06.011
- Wang, Q., Garrity, G.M., Tiedje, J.M., Cole, J.R., 2007. Naive Bayesian Classifier for Rapid Assignment of rRNA Sequences into the New Bacterial Taxonomy. *Appl. Environ. Microbiol.* 73, 5261–5267. <https://doi.org/10.1128/AEM.00062-07>
- Wangenstein, O. S., Palacín, C., Guardiola, M., & Turon, X. (2018). DNA metabarcoding of littoral hard-bottom communities: high diversity and database gaps revealed by two molecular markers. *PeerJ*, 6, e4705. doi: 10.7717/peerj.4705
- Wedding, L. M. M., Friedlander, A. M. M., Kittinger, J. N. N., Watling, L., Gaines, S. D. D., Bennett, M., ... Smith, C. R. R. (2013). From principles to practice: a spatial approach to systematic conservation planning in the deep sea. *Proceedings of the Royal Society B*, 280(1773), 20131684. doi: 10.1098/rspb.2013.1684
- Wedding, L. M. M., Reiter, S. M., Smith, C. R., Gjerde, K. M., Kittinger, J. N., Friedlander, A. M., ... Crowder, L. B. (2015). Managing mining of the deep seabed. *Science*, 349(6244), 144–145. doi: 10.1126/science.aac6647
- White, M., Bashmachnikov, I., Arstegui, J., & Martins, A. (2008). Physical Processes and Seamount Productivity. In *Seamounts: Ecology, Fisheries & Conservation* (pp. 62–84). doi: 10.1002/9780470691953.ch4
- Wilkinson, S., Davy, S., Bunce, M., & Stat, M. (2018). Taxonomic identification of environmental DNA with informatic sequence classification trees. *PeerJ Preprints*, (March), 1–2. doi: 10.7287/peerj.preprints.26812
- Wilson, R. R., & Kaufmann, R. S. (1987). Seamount Biota and Biogeography. In B. H. Keating, P. Fryer, R. Batiza, & G. W. Boehlert (Eds.), *Seamounts, Islands, and Atolls* (pp. 355–377). doi: 10.1029/GM043p0355
- Wu, S., Xiong, J., & Yu, Y. (2015). Taxonomic Resolutions Based on 18S rRNA Genes: A Case Study of Subclass Copepoda. *PLoS ONE*, 10(6), e0131498. doi: 10.1371/journal.pone.0131498

Accepted Article

Zeppilli, D., Bongiorni, L., Cattaneo, A., Danovaro, R., & Santos, R. S. (2013). Meiofauna assemblages of the Condor Seamount (North-East Atlantic Ocean) and adjacent deep-sea sediments. *Deep-Sea Research Part II: Topical Studies in Oceanography*, 98(PA), 87–100. <https://doi.org/10.1016/j.dsr2.2013.08.009>

Zeppilli, D., Bongiorni, L., Santos, R. S., & Vanreusel, A. (2014). Changes in nematode communities in different physiographic sites of the Condor Seamount (North-East Atlantic Ocean) and adjacent sediments. *PLoS ONE*, 9(12), 1–26. <https://doi.org/10.1371/journal.pone.0115601>

Zhan, A., Hulák, M., Sylvester, F., Huang, X., Adebayo, A. A., Abbott, C. L., ... Macisaac, H. J. (2013). High sensitivity of 454 pyrosequencing for detection of rare species in aquatic communities. *Methods in Ecology and Evolution*, 4(6), 558–565. doi: 10.1111/2041-210X.12037

Tables and figures

Tables

Table 1. Mean percent of metazoan amplicon sequence variant (ASV; 18S) and operational taxonomic unit (OTU; COI) that could be assigned taxonomy at each level.

Target gene	Phylum	Class	Order	Family	Genus	Species
18S rRNA	86.7	77.43	69.19	24.96	19.56	17.79
COI	18.59	7.68	3.21	1.48	0.88	0.69

Table 2. Mean of estimated particulate organic carbon (POC) flux ($\text{gC m}^{-2} \text{yr}^{-1}$) and polymetallic nodule abundance (kg m^{-2}) at our sampling sites within each APEI. Estimates of POC flux derive from the global model reported in Lutz et al. (2007), and nodule abundance from the geological model described in ISA Technical Study No. 6 (*also see* Table S14). APEI = Area of Particular Environmental Interest.

APEI	POC flux		Nodule abundance	
	Mean	SD	Mean	SD
1	1.13	0.025	2.14	0.311
4	1.40	0.041	5.83	0.089
7	1.88	0.060	0.58	0.002

Figures

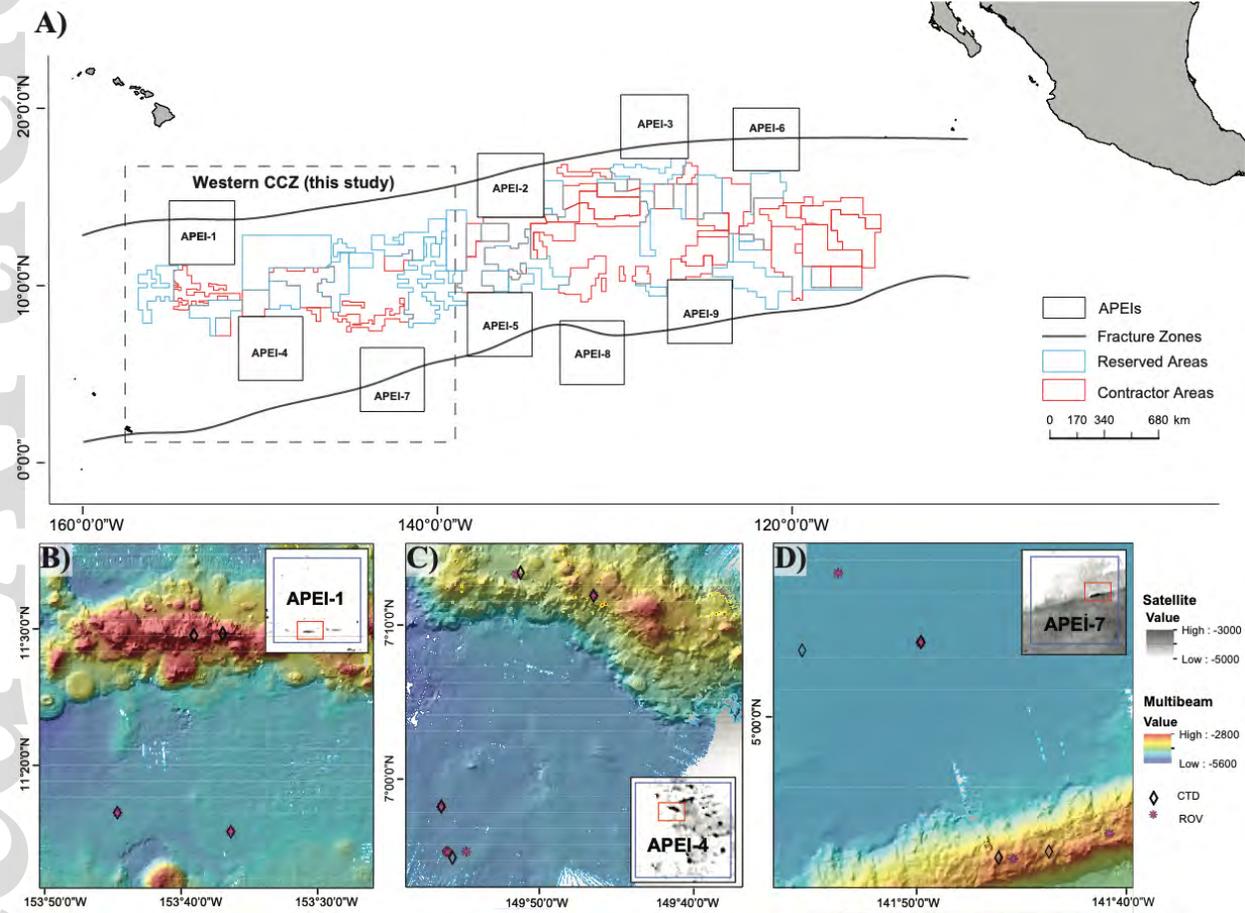


Fig. 1. Maps of the study areas within the Clarion Clipperton Zone. (A) Overview of the CCZ and location of the APEIs. Sampling locations within APEI 1 (B), APEI 4 (C) and APEI 7 (D), with symbols for collection types and inset map of the seamount location within the APEI. APEI = Area of Particular Environmental Interest, designated as no-mining areas by the ISA.

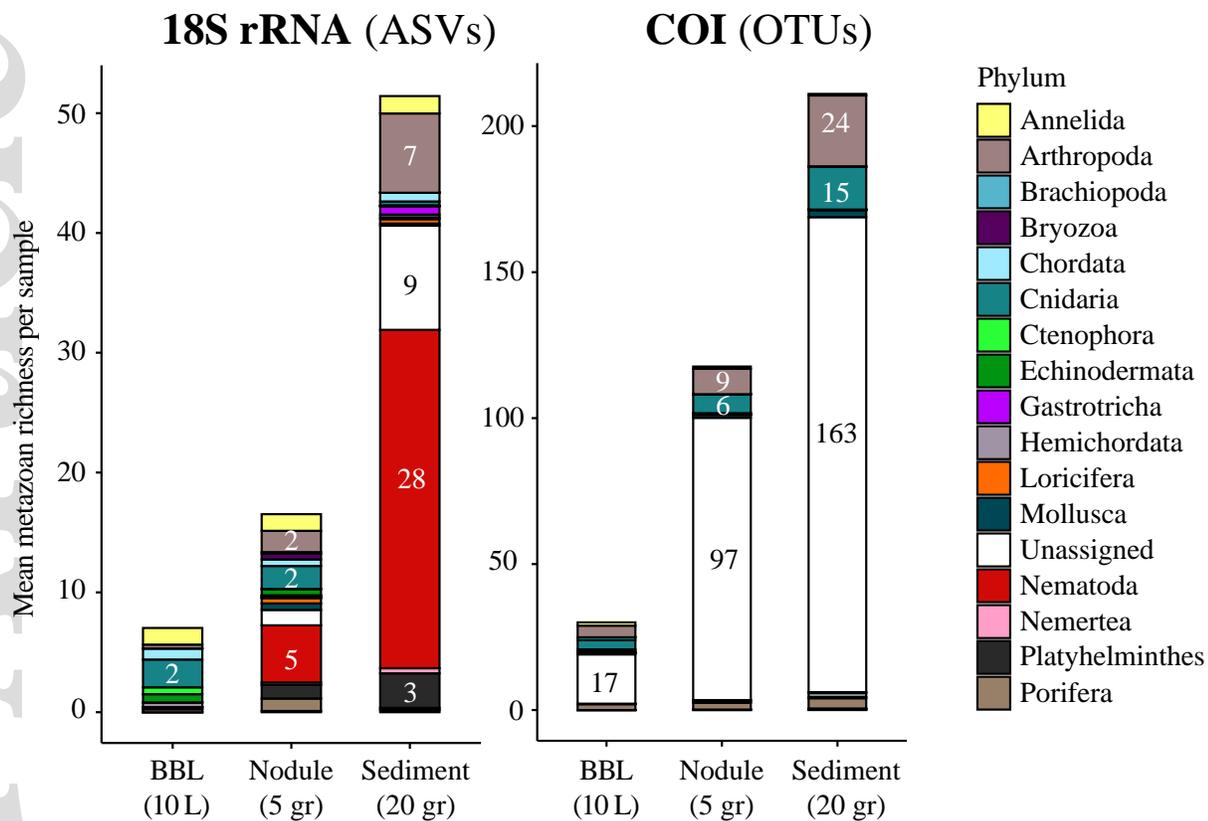


Fig. 2. Barplots of mean metazoan 18S amplicon sequence variants (ASVs) and COI operational taxonomic units (OTUs), shown per sample type and coloured by Phylum. BBL = Benthic boundary layer. Numbers inside the histogram bars correspond to mean number of ASVs per phylum. Only the ten most abundant phyla for 18S and COI are shown.

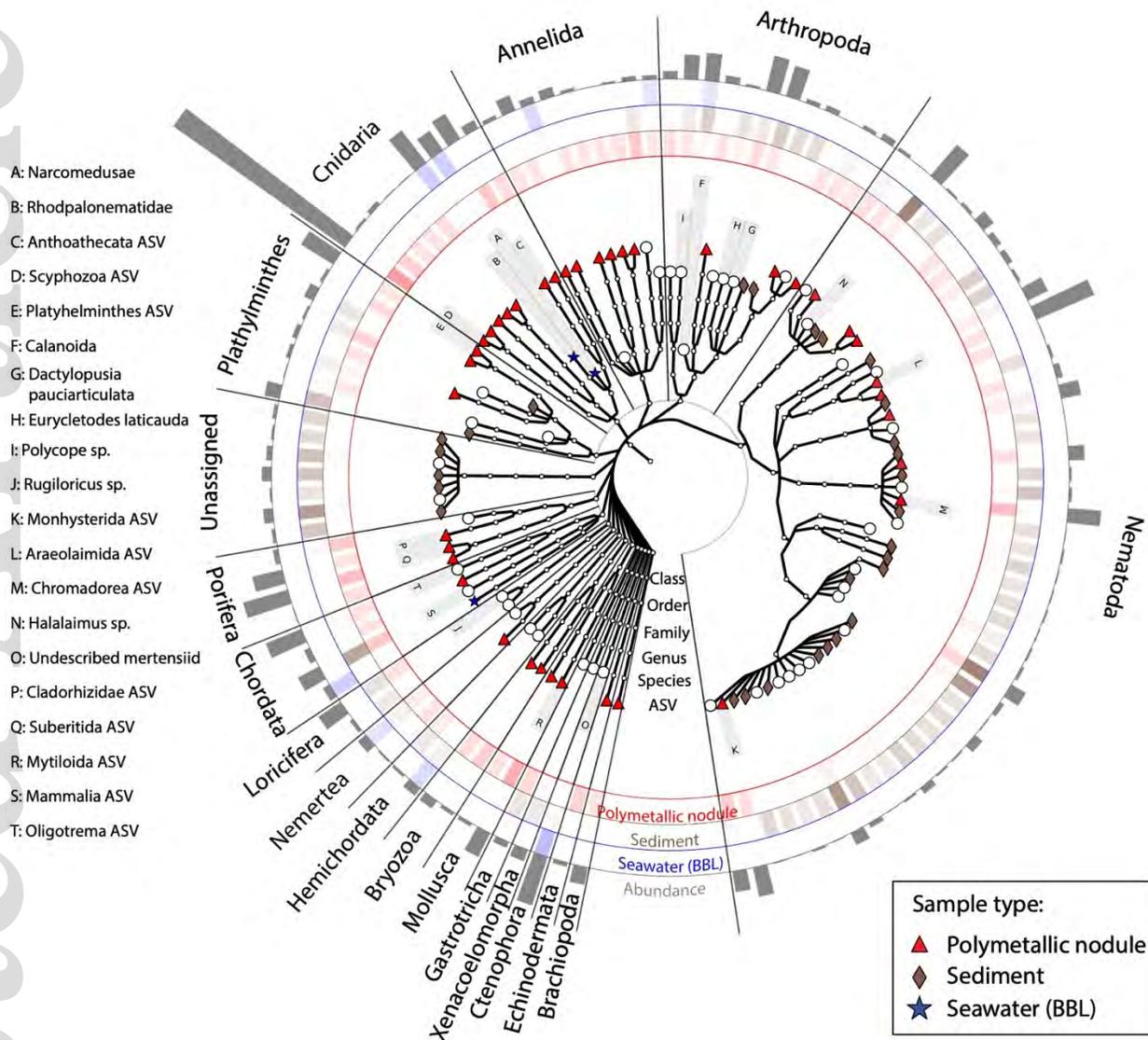


Fig. 3. Cladogram with circular heatmap and barplots for the metazoan community resolved by 18S rRNA. The colour intensity in the circular heatmap corresponds to mean relative abundance in each sample type across the whole dataset. The bar heights on the outside of the circle are proportional to the mean relative abundance of each taxon within the entire dataset. Taxa found exclusively in one sample type are marked by a corresponding symbol: red triangle for nodules, grey diamond for sediment and blue star for BBL. Those found in more than one sample type are marked by a white circle. The 20 most abundant taxa at the tip of each branch are labelled with letters, and identified to highest taxonomic resolution (key at left). Only taxa found in a minimum of five samples were included. BBL = benthic boundary layer.

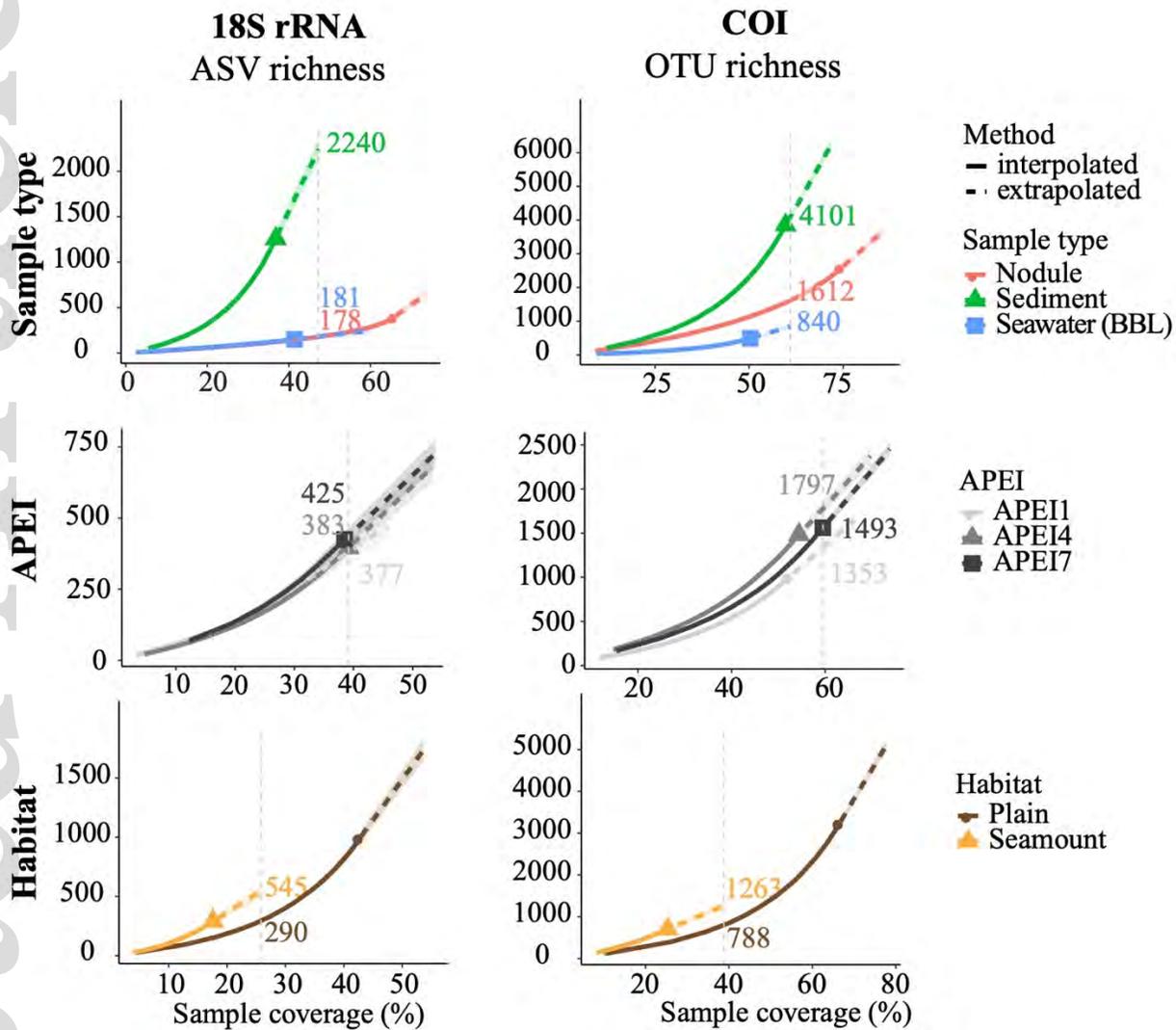


Fig. 4. Metazoan 18S amplicon sequence variant (ASVs) and COI operational taxonomic unit (OTU) gamma diversity per APEI and habitat variable at base sampling coverage. ASV and OTU richness were estimated using Chao2. Shaded coloured areas indicate the 95 % confidence intervals obtained using a bootstrap method with 200 replicates. Coloured numbers in the plots represent number of ASVs/OTUs at base coverage. BBL = benthic boundary layer. For APEIs and habitats comparisons, only sediment samples were included. Additionally, for the APEI comparison, seamount samples were excluded, as not all APEIs had seamount sediment data.

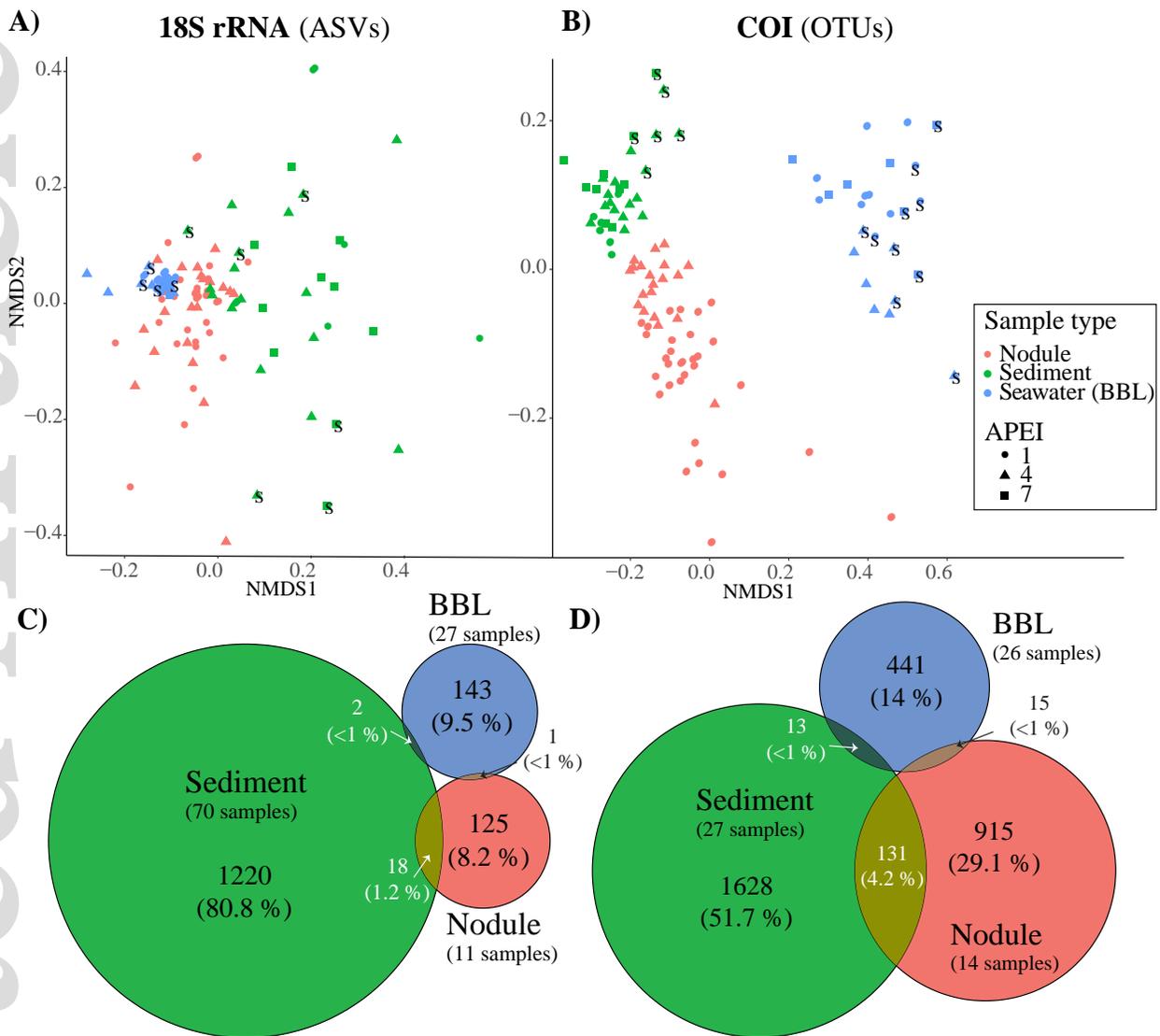


Fig. 5. Community similarity across sample/substrate type and habitat. (A, B) Non-metric multi-dimensional scaling plots (nMDS) of metazoan community dissimilarity, and (C, D) Venn diagrams illustrating shared metazoan amplicon sequence variant (ASVs; 18S) and operational taxonomic units (OTUs; COI) between sample types. nMDS plots are based on dissimilarity matrices using unweighted unifracs distance. Seamount samples in A-B are indicated by the letter 'S'. Results in C-D represent mean values of 50 subsampling iterations. Subsampling was performed to normalise the number of samples per sample type at equivalent sampling coverage (coverage of 40 % and 50 % for 18S and COI, respectively), estimated using the Chao2 index. BBL = benthic boundary layer, APEI =

area of principal environmental interest, ASV = amplicon sequence variant, OTU = operational taxonomic unit.

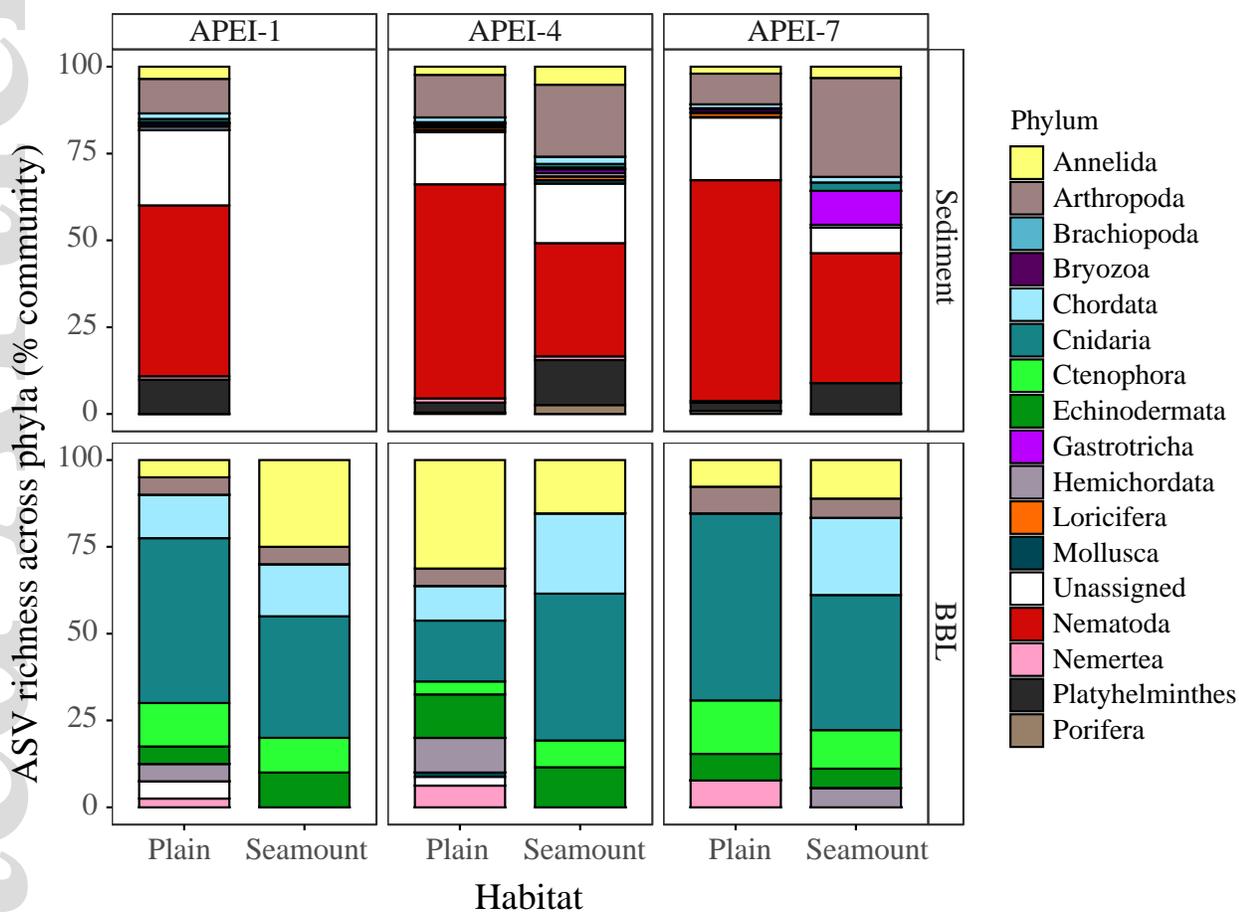


Fig. 6. Community composition on seamounts and abyssal plains for each APEI. Relative 18S amplicon sequence variant (ASV) richness across phyla per sample, per habitat and per APEI. BBL = Benthic boundary layer, indicating seawater sampled within the BBL.

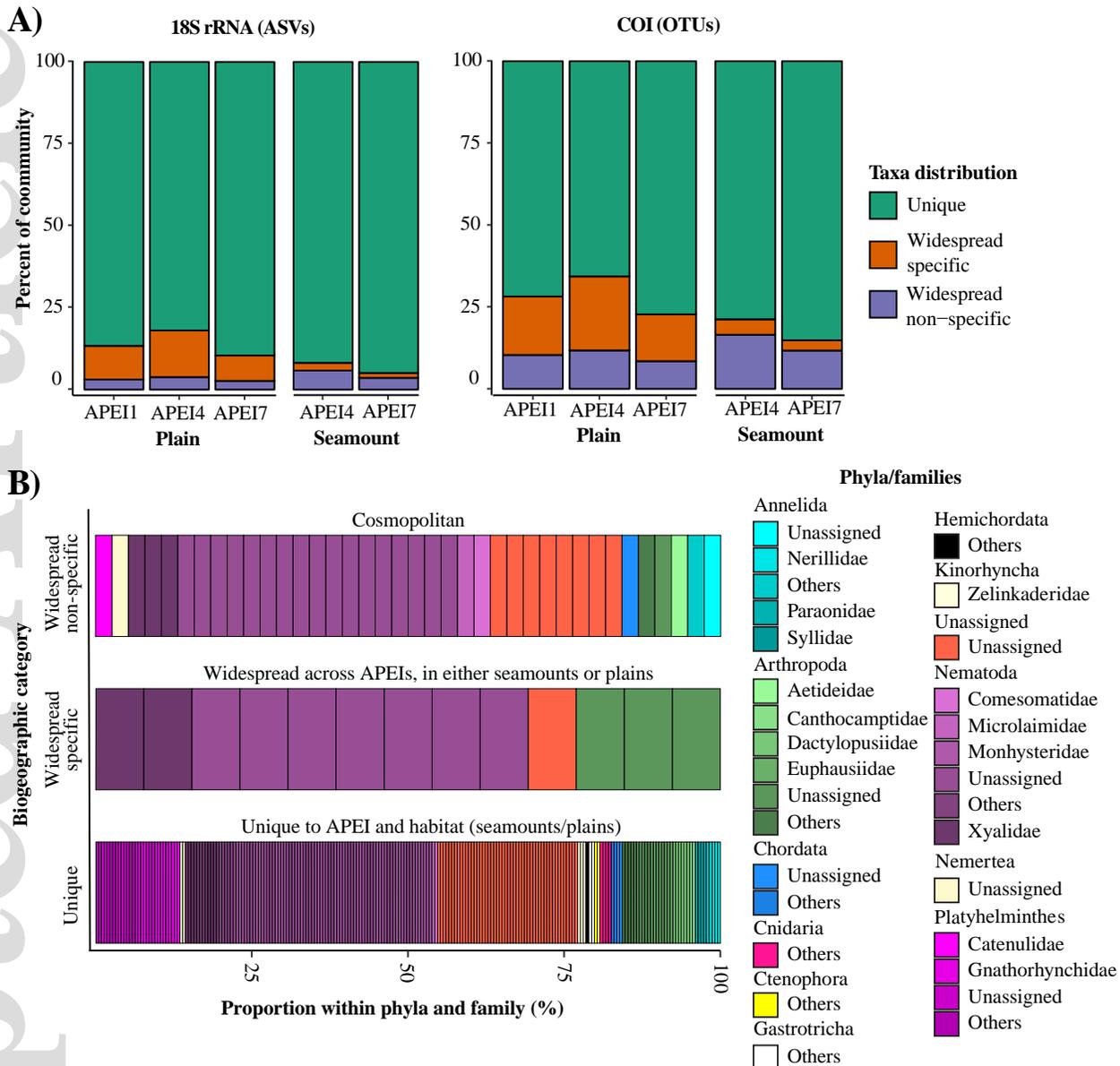


Fig. 7. Community biogeography of abyssal seamounts and plains. (A) Proportion of sediment 18S amplicon sequence variants (ASVs) and COI operational taxonomic units (OTUs) found to be either unique to each APEI and habitat combination (Unique), found within more than one APEI but only one habitat (widespread-specific), or found within more than one APEI and habitat (widespread non-specific). (B) Taxonomic information for 18S ASVs within each biogeographic category. ASVs in 7B are delimited by thin black lines. In 7B, results from all iterations were used to assign ASVs to taxonomic groups.